

**Impact of BVOC on
tropospheric
chemistry**

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Impact of biogenic hydrocarbons on tropospheric chemistry: results from a global chemistry-climate model

G. A. Folberth^{1,*}, D. A. Hauglustaine¹, J. Lathière¹, and F. Brocheton²

¹Laboratoire des Sciences du Climat et de l'Environnement (LSCE), Gif-sur-Yvette, France

²Centre National de Recherches Météorologiques (CNRM), Météo France, Toulouse, France

*now at: School of Earth and Ocean Science (SEOS), University of Victoria, Victoria, Canada

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Correspondence to: G. A. Folberth (gerd.folberth@ec.gc.ca)

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Abstract

We present a description and evaluation of LMDz-INCA, a global three-dimensional chemistry-climate model, pertaining to its recently developed NMHC version. In this substantially extended version of the model a comprehensive representation of the photochemistry of non-methane hydrocarbons (NMHC) and volatile organic compounds (VOC) from biogenic, anthropogenic, and biomass-burning sources has been included. The tropospheric annual mean methane (9.2 years) and methylchloroform (5.5 years) chemical lifetimes are well within the range of previous modelling studies and are in excellent agreement with estimates established by means of global observations. The model provides a reasonable simulation of the horizontal and vertical distribution and seasonal cycle of CO and key non-methane VOC, such as acetone, methanol, and formaldehyde as compared to observational data from several ground stations and aircraft campaigns. LMDz-INCA in the NMHC version reproduces tropospheric ozone concentrations fairly well throughout most of the troposphere. The model is applied in several sensitivity studies of the biosphere-atmosphere photochemical feedback. The impact of surface emissions of isoprene, acetone, and methanol is studied. These experiments show a substantial impact of isoprene on tropospheric ozone and carbon monoxide concentrations revealing an increase in surface O₃ and CO levels of up to 30 ppbv and 60 ppbv, respectively. Isoprene also appears to significantly impact the global OH distribution resulting in a decrease of the global mean tropospheric OH concentration by approximately 0.9×10^5 molecules cm⁻³ or roughly 10% and an increase in the global mean tropospheric methane lifetime by approximately four months. A global mean ozone net radiative forcing due to the isoprene induced increase in the tropospheric ozone burden of 0.09 W m^{-2} is found. The key role of isoprene photooxidation in the global tropospheric redistribution of NO_x is demonstrated. LMDz-INCA calculates an increase of PAN surface mixing ratios ranging from 75 to 750 pptv and 10 to 250 pptv during northern hemispheric summer and winter, respectively. Acetone and methanol are found to play a significant role in the up-

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per troposphere/lower stratosphere (UT/LS) budget of peroxy radicals. Calculations with LMDz-INCA show an increase in HO_x concentrations region of 8 to 15% and 10 to 15% due to methanol and acetone biogenic surface emissions, respectively. The model has been used to estimate the global tropospheric CO budget. A global CO source of 3019 TgCO yr⁻¹ is estimated. This source divides into a primary source of 1533 TgCO yr⁻¹ and secondary source of 1489 TgCO yr⁻¹ deriving from VOC photooxidation. Global VOC-to-CO conversion efficiencies of 90% for methane and between 20 and 45% for individual VOC are calculated by LMDz-INCA.

1. Introduction

Non-Methane volatile organic compounds (NMVOC) are known to affect the chemical composition of the atmosphere decisively. NMVOC play a key role in the sequestration of nitrogen oxides (NO_x) via the formation of organic nitrates (e.g., PAN and analogs), can directly or indirectly increase the acidity of precepitation, and provide the starting material for much of the natural atmospheric aerosols. They have been found to significantly contribute to the production of pollutants and are of major concern in the assessment and controlling strategies of present-day and future air quality (e.g. Graedel, 1979; Brasseur and Chatfield, 1991; Crutzen and Zimmermann, 1991; Fehsenfeld et al., 1992; Andreae, 1995; Crutzen, 1995; Andreae and Crutzen, 1997; Berntsen et al., 1997; Levy II et al., 1997; Wang et al., 1998c; Granier et al., 2000; Hauglustaine and Brasseur, 2001). But most important, NMVOC play a central role in tropospheric ozone formation.

Ozone is a key component in the atmosphere. It is an effective oxidant and greenhouse gas, especially in the upper troposphere (Lacis et al., 1990; Hauglustaine et al., 1994). In addition, near the surface ozone can have detrimental effects on the vegetation and on human health (Fishman, 1991; Finlaysonpitts and Pitts, 1993; Taylor, 2001; Bernard et al., 2001). Ozone photolysis by ultraviolet radiation is the primary source of hydroxyl radicals in the troposphere. Photochemical oxidation of NMVOC, on the other

hand, is primarily initiated and, hence, controlled by reaction with OH. This reaction determines the magnitude and distribution of hydroxyl radical concentrations, thereby altering the oxidative capacity of the troposphere (Houweling et al., 1998; Wang et al., 1998c; Poisson et al., 2000).

5 The direct radiative forcing due to NMVOC has been found to be negligibly small. An upper limit to the global mean anthropogenic forcing of 0.015 W m^{-2} has been established by Highwood et al. (1999). Collins et al. (2002), on the other hand, have presented a study, which convincingly demonstrates that NMVOC are able to exert a substantial indirect effect on greenhouse warming by affecting ozone formation and the methane lifetime towards reaction with OH. Moreover, the formation of secondary organic aerosols (SOA) in the course of photochemical NMVOC oxidation is believed to have a direct and indirect effect on the radiative flux of the lower atmosphere (Kanaki-
10 dou et al., 2000; Tsigaridis and Kanakidou, 2003).

NMVOC primarily originate from three principle sources: anthropogenic activities, biomass burning, and the biosphere. The biosphere acts as the largest source of reactive trace gases in the troposphere. It has been suggested that the biogenic source on the global scale surpasses several times the combined NMVOC emission flux originating from anthropogenic and biomass burning sources. State-of-the-art emission estimates include an annual global BVOC source of approximately 750 TgC yr^{-1} (Guenther
15 et al., 1995) whereas the anthropogenic and biomass burning sources together amount to roughly 90 TgC yr^{-1} of NMVOC (Hao and Liu, 1994; Olivier et al., 1996, 2001; Olivier and Berdowski, 2001; Van der Werf et al., 2003).

Biogenic volatile organic compounds (BVOC) include isoprene and isoprenoid compounds (such as monoterpenes and higher terpenes) as well as a large number of other species from the groups of alkanes, non-isoprenoid alkenes, carbonyls, alcohols, and organic acids. They are emitted into the atmosphere from natural sources in terrestrial and marine ecosystems. In terms of abundance and importance the predominant BVOC are isoprene and terpenes, methanol, and acetone (Bonsang et al., 1992; Mac-
25 Donald and Fall, 1993; Sharkey and Singsaas, 1995; Kirstine et al., 1998; Bonsang

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and Boissard, 1999; Doskey and Gao, 1999; Guenther et al., 2000; Singh et al., 2000; Galbally and Kirstine, 2002; Jacob et al., 2002).

An increasing importance of isoprene (Shallcross and Monks, 2000) and other BVOC (Guenther et al., 1999; Kellomaki et al., 2001) in the future has been hypothesized due to an increasing net primary production associated with a warmer climate (Constable et al., 1999). If global patterns and magnitudes of biogenic VOC emissions change in correlation with climate-related alterations in temperature, precipitation, and solar insolation, in turn a feed back upon the climate via changes in the accumulation rate of atmospheric greenhouse gases seems very likely.

Hauglustaine et al. (2004) recently presented the global climate-chemistry model LMDz-INCA, which takes into account the CH₄-NO_x-CO-O₃ chemistry of the background troposphere. This model has been supplemented by a detailed non-methane hydrocarbon scheme in order to investigate biosphere-atmosphere interactions. This work represents a further step in the framework of several ongoing studies (Boucher et al., 2002; Hauglustaine et al., 2004; Bauer et al., 2004), which eventually will converge towards a modelling system that takes into account the “complete” chemistry of the troposphere and stratosphere, including the different types of aerosols, in a fully interactive Earth System Model.

In this work we present the non-methane hydrocarbon version of INCA (version NMHC.1.0). A description and general evaluation of this new model version is provided. The more general application abilities are than used to investigate various aspects of biosphere NMVOC emissions and tropospheric chemical composition including ozone formation, tropospheric HO_x, and possible impacts on future climate.

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2. Model description

2.1. The LMDz General Circulation Model

LMDz (Laboratoire de Météorologie Dynamique, zoom) is a grid point General Circulation Model (GCM) developed initially for climate studies by [Sadourny and Laval \(1984\)](#).
5 As part of the IPSL Earth System Model, the GCM lately has undergone a major re-coding and has been applied in climate feedback studies by [Friedlingstein et al. \(2001\)](#) and [Dufresne et al. \(2002\)](#).

In LMDz the finite volume transport scheme of [Van Leer \(1977\)](#) as described in [Hourdin and Armengaud \(1999\)](#) is used to calculate large-scale advection of tracers. The
10 parametrization of deep convection is based on the scheme of [Tiedtke \(1989\)](#); a local second-order closure formalism is used to describe turbulent mixing in the planetary boundary layer (PBL).

The LMDz version used in this study (referred to as 3.3) has a horizontal resolution of 3.8 degrees in longitude and 2.5 degrees in latitude (96×72 grid cells). The model is
15 composed of 19 vertical levels on σ -p coordinates extending from the surface to 3 hPa. Also higher-resolution versions of LMDz have been developed and applied recently (e.g. [Bauer et al., 2004](#)).

The primitive equations in the GCM are solved with a 3 min timestep, large-scale transport of tracers is carried out every 15 min, and physical processes are calculated
20 at a 30 min time interval. For a more detailed description and an extended evaluation of the GCM we refer to the work of, e.g., [Le Treut et al. \(1994\)](#) and [Harzallah and Sadourny \(1995\)](#). Recently, [Hauglustaine et al. \(2004\)](#) have coupled LMDz to the tropospheric chemistry model INCA, during which the GCM has been reevaluated.

2.2. The INCA-NMHC Chemistry and Aerosol Model

25 INteraction with Chemistry and Aerosols (INCA) is coupled on-line to the LMDz general circulation model. INCA prepares the surface and in situ emissions, calculates dry

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deposition and wet scavenging rates, and integrates in time the concentration of atmospheric species with a time step of 30 min. INCA uses a sequential operator approach, a method generally applied in chemistry-transport-models (Muller and Brasseur, 1995; Brasseur et al., 1998; Wang et al., 1998a; Poisson et al., 2000). The INCA-NMHC version applied in this study is based on an earlier version which was developed to represent the background chemistry of the troposphere (Hauglustaine et al., 2004). Results from this version of the model related to the impact of chemistry on the budget of CO₂ have been published by Folberth et al. (2005).

2.2.1. Chemistry

The version of INCA used in this study includes a comprehensive photochemical scheme originally intended to represent the photochemistry of the troposphere in regional scale chemistry-transport models. In addition to the CH₄-NO_x-CO-O₃ photochemistry representative of the tropospheric background, INCA, in its NMHC-implementation, also takes into account the photochemical oxidation pathways of non-methane hydrocarbons (NMHC) and non-methane volatile organic compounds (NMVOC) from natural and anthropogenic sources as well as their photochemical oxidation products. The model can be applied to calculate the distribution of tropospheric ozone and its precursors, but due to the comprehensive chemistry scheme the model is also suited to be used in studies of biosphere-atmosphere interrelation and the impact of a changing biosphere on the global climate. The species included in this more extensive version of LMDz-INCA are summarized in Table 1. We consider only one photochemical family (O_x = O₃ + O(¹D) + O(³P)), and for species with very short photochemical lifetimes transport is not taken into account.

Tables 1, 2, 3, and 4 summarize the chemical scheme in LMDz-INCA. The scheme includes a total of 83 species, 58 of which are subject to transport. In INCA-NMHC short-chained NMVOC are treated explicitly whereas a lumping approach is applied in the case of higher NMVOC as proposed by Brocheton (1999). Alkanes up to three carbon atoms (C₃) per molecule are treated explicitly including methane, ethane, and

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propane. All C₄- and higher alkanes are lumped into one artificial species (C₄⁺-alkanes). Likewise, the model distinguishes ethene, propene, and C₄⁺-alkenes. The isoprene oxidation pathway has been represented with some complexity, comparable to the one proposed by [Poschl et al. \(2000\)](#) including explicitly the major oxidation products methyl vinyl ketone and methacrolein; all higher isoprenoid species are lumped into one group to which we refer to as “terpenes”. INCA-NMHC includes two alcohols (methanol and C₂⁺-alcohols, eleven hydroperoxides, one group representing organic acids, three aldehyde species (formaldehyde, acetaldehyde including higher mono-carboxy aldehydes, and a group representing higher aldehydes produced by terpene oxidation), three ketones (acetone, methyl ethyl ketone, and methyl vinyl ketone), four species representing peroxy acetyl nitrate (PAN) and analogues, two general organic nitrate groups (organic nitrates with low or high reactivity towards OH), as well as the corresponding organic radicals arising in the oxidation of the above species (cf. Tables 1 and 4). Lumping in all cases follows the generally applied method of grouping individual compounds by their reactivity towards the OH-radical while also taking into account compound groups, molecular weights, and atmospheric abundances (for a general description of these methods see, e.g., [Stockwell et al., 1997](#)).

43 photolytic reactions (Table 2), 217 thermochemical reactions (Table 4), and 4 heterogeneous reactions (Table 3) are taken into account by the model. The set of reactions and reaction rates is based on the scheme of [Brocheton \(1999\)](#). All reaction rates have been reviewed and updated with respect to the compilations of [Atkinson et al. \(1997\)](#) and [DeMore et al. \(1997\)](#) as well as subsequent updates ([Sander et al., 2000, 2002](#)). The reaction rates are calculated by the model at each time interval on the basis of temperature, pressure, and water vapor distribution provided by the GCM.

Pretabulated clear-sky photolysis frequencies *j* are used to determine the values at each time step and model grid cell by means of a multivariate log-linear interpolation through a Taylor series expansion ([Burden and Faires, 1985](#)). This look-up table is prepared with the Troposphere-Ultraviolet-Visible (TUV) model (version 4.1) from [Madronich and Flocke \(1998\)](#) using a pseudo-spherical 16 stream discrete-ordinate

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method. The effect of cloudiness is taken into account according to [Chang et al. \(1987\)](#) as described by [Brasseur et al. \(1998\)](#). Further details can be found in [Hauglustaine et al. \(2004\)](#). Where no spectral data is available for a specific species to be used with TUV we assume the photolysis frequencies to be linearly dependent on chemically similar compounds (cf. Table 2). The LMDz-INCA chemical scheme includes four heterogeneous reactions (Table 3) following the recommendations of [Jacob \(2000\)](#) and using the monthly averaged sulfate aerosol fields from [Boucher et al. \(2002\)](#) to determine the aerosol distribution as discussed by [Hauglustaine et al. \(2004\)](#). The NMHC-version of LMDz-INCA uses the identical parameterization for the representation of these processes.

The mechanism implicitly accounts for a carbon loss through formation of secondary organic aerosols (SOA). This tropospheric carbon sink is taken into account in the scheme as a carbon imbalance for specific reactions in the terpene oxidation pathway. The estimate of the global carbon sink as derived with LMDz-INCA amounts to approximately 39% ($\sim 37 \text{ TgC yr}^{-1}$) of the total annual carbon surface flux emitted as terpenes. Our estimate seems in satisfactory agreement with the calculated variation of the global annual SOA production of 2.5 to 44 TgC yr^{-1} for the biogenically produced SOA ([Tsigaridis and Kanakidou, 2003](#)). Note that explicit secondary organic aerosol formation is not calculated in the current implementation of LMDz-INCA and SOA are not included in the INCA species inventory.

Interactive calculation of chemistry and transport of species extends up to the upper model level. However, in the current version of INCA no chlorine or bromine chemistry is taken into account and heterogeneous reactions on polar stratospheric clouds (PCS) is not yet considered. Therefore, ozone concentrations are relaxed toward observations at the uppermost model levels at each time step applying a relaxation time constant of ten days above the 380 K potential temperature surface. The ozone observations are taken from the monthly mean 3-D climatologies of [Li and Shine \(1995\)](#). Further details and an evaluation of the stratospheric boundary conditions in INCA can be found in [Hauglustaine et al. \(2004\)](#); the NMHC version of INCA applies the same

approach.

Forward integration in time of the chemical equations is conducted on the basis of five possible numerical algorithms with a time step of 30 min as described by [Hauglustaine et al. \(2004\)](#). In the current configuration of LMDz-INCA we apply the explicit Euler forward algorithm ([Brasseur et al., 1999](#)) for the integration of long-lived species and the implicit Euler backward with Newton-Raphson iteration scheme for all other species (cf. Table 1). A species to be long-lived in the above sense and suitable for the less resourceful numerical algorithm requires the mean atmospheric lifetime of this species to be at least a few months.

2.2.2. Emissions

Surface emission inventories have been compiled for this version of LMDz-INCA using the same methods of data preparation as described in [Hauglustaine et al. \(2004\)](#). These inventories are based on various compilations, with the exception of the biogenic surface source. Biogenic surface emissions were prepared using a global vegetation model which includes a biogenic emission module. Table 5 summarizes the emission magnitude for the individual species that are considered in the version of LMDz-INCA and used in the current study.

LMDz-INCA uses anthropogenic emissions (industry, fossil fuel, and industrial bio-fuel) based on the EDGAR v3.2 and EDGAR v2.0 emission databases. Anthropogenic sources for nitrogen oxides (NO_x), carbon monoxide (CO), and methane (CH_4) are introduced on the basis of the estimates provided in EDGAR v3.2 ([Olivier and Berdowski, 2001](#)) representative of the year 1995. The CO global emissions are rescaled to the values given by [Prather et al. \(2001\)](#). In addition, the LMDz-INCA emission inventory also includes NO_x emissions from oceangoing ships based on [Corbett et al. \(1999\)](#) and NO_x aircraft emissions based on the ANCAT/EC2 inventory ([Gardner et al., 1998](#)).

At the time of creation of the current LMDz-INCA emission inventory the more recent version EDGAR v3.2 did not yet include NMVOC emissions on a per-compound basis. Therefore, it was decided to use anthropogenic NMVOC emission estimates as

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given by the EDGAR v2.0 database (Olivier et al., 1996) which includes estimates for individual compounds. NMVOC emissions based on EDGAR v2.0 are representative of the year 1990 and include alkanes (C_2H_6 , C_3H_8 , C_4^+ -alkanes), alkenes and alkynes (C_2H_4 , C_3H_6 , C_4^+ -alkenes, C_2H_2), aromatics, aldehydes (CH_2O , CH_3CHO plus higher aldehydes) as well as ketones (CH_3COCH_3 and higher ketones, methylethyl ketone, methylvinyl ketone). Species explicitly included in EDGAR v2.0 that are not explicit in LMDz-INCA were aggregated into the specific compound group based on their individual molecular weights, their reactivity towards the hydroxyl radical, and their relative atmospheric abundances.

Biomass burning emissions are introduced according to the satellite based inventory developed by Van der Werf et al. (2003), averaged over the period 1997–2001. This data set is used to provide the geographical distribution and seasonal variation for each compound. Domestic biofuel use and agricultural waste burning emissions are also included in the biomass burning category and are based on the EDGAR database. Emission factors compiled by Andreae and Merlet (2001) are then used to derive the emission magnitude of the biomass burning surface flux for the individual compounds.

NO soil emissions are introduced on the basis of Yienger and Levy II (1995), emissions of CH_4 from rice paddies, wetlands, termites, wild animal and ruminants, and the ocean are taken from Fung et al. (1991), and emissions of N_2O are based on Bouwman and Taylor (1996) and Kroeze et al. (1999) for continental emissions and on Nevison and Weiss (1995) for the oceanic N_2O surface flux. NO emissions from lightning are calculated interactively in LMDz-INCA on the basis of the occurrence of convection and cloud top heights. The parameterization of NO from lightning sources as it is used in LMDz-INCA is discussed in more detail by Jourdain and Hauglustaine (2001) and Hauglustaine et al. (2004).

Biogenic emissions of isoprene, terpenes, methanol, and acetone have been prepared with the dynamical global vegetation model ORCHIDEE (Organizing Carbon and Hydrology in Dynamic EcosystEms) (Krinner et al., 2005). ORCHIDEE essentially includes three different components, namely the surface-vegetation-atmosphere transfer

scheme SECHIBA (Ducoudré-De Noblet et al., 1993; de Rosnay and Polcher, 1998), the dynamic global vegetation model LPJ (Sitch et al., 2003), and STOMATE (Saclay-Toulouse-Orsay Model for the Analysis of Terrestrial Ecosystems), a newly developed model simulating plant phenology and carbon dynamics. For a detailed description and extended evaluation of ORCHIDEE and the new carbon dynamics model STOMATE we refer to the paper by Krinner et al. (2005).

In order to calculate NMVOC surface fluxes from the terrestrial biosphere, a biogenic emission module has recently been integrated into ORCHIDEE. The calculation is based on the emission model by Guenther et al. (1995) and uses input from ORCHIDEE for the key parameters (leaf area index, PFT distribution, specific leaf weight, etc.). This model takes into account changes in the flux strength due to leaf temperature, photosynthetically active radiation (direct and diffuse), and leaf aging. A detailed description and evaluation of the terrestrial NMVOC emission model is given by Lathière et al. (2005).

Warneke et al. (1999) first suggested a significant source of methanol from decaying plant matter and recent estimates of its magnitude range between 4 and 17 TgC yr⁻¹ (Singh et al., 2000; Heikes et al., 2002; Galbally and Kirstine, 2002). The LMDZ-INCA emission inventory includes biogenic methanol emissions deriving from plant decay with a global source strength of 13 TgC yr⁻¹. It is assumed that this methanol source collocates with methanol emissions from the terrestrial vegetation.

LMDz-INCA takes into account oceanic emissions of CO and several NMVOC (cf. Table 5 for compounds that possess oceanic sources). The spatial distribution and variation in time of oceanic CO emissions is based on Erickson and Taylor (1992) and scaled to a global mean of 50 TgC yr⁻¹ (Prather et al., 2001). It was furthermore assumed that the global geographic distribution of oceanic NMVOC emissions equals the global oceanic CO source distribution. The oceanic NMVOC emission magnitude is based on Jacob et al. (2002) in case of acetone and on the work of Bonsang et al. (1992) and Bonsang and Boissard (1999) for all other NMVOC that possess non-zero oceanic sources in the LMDz-INCA emission inventory.

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2.3. Dry deposition and wet removal

Dry deposition in LMDz-INCA is based on the resistance-in-series approach (Wesely, 1989; Walmsley and Wesely, 1996; Wesely and Hicks, 2000). Deposition velocities (v_d) are calculated at each time step according to:

$$v_d = \frac{1}{R_a + R_b + R_c}, \quad (1)$$

where R_a , R_b , and R_c (s/m) are the aerodynamical, quasi-laminar, and surface resistance, respectively. R_a and R_b are determined on the basis of Walcek et al. (1986). The surface resistance calculation for all species included in LMDz-INCA is based on their temperature dependent Henry's Law Equilibrium Constant and reactivity factor for the oxidation of biological substances. Henry's Law Coefficients tabulated for standard conditions have been taken from Sander (1999) and reactivity factors are taken from Wesely (1989) and Walmsley and Wesely (1996). The vegetation map classification of De Fries and Townshend (1994), interpolated to the model grid and redistributed into the classification of Wesely (1989), is used to parameterize land use dependencies of the surface resistance R_c . A complete list of species that are subject to dry deposition is given in Table 1.

Wet scavenging in INCA is parameterized as a first-order loss process as proposed by Giorgi and Chameides (1985):

$$\frac{d}{dt}C_g = -\beta C_g, \quad (2)$$

where C_g is the gas-phase concentration of the considered species and β is the scavenging coefficient (1/s). Wet scavenging associated with large scale stratiform precipitation is calculated adopting the falling raindrop approach (Seinfeld and Pandis, 1998) and wet removal of soluble species by convective precipitation is calculated as part of the upward convective mass flux on the basis of a modified version of the scheme proposed by Balkanski et al. (1993). INCA calculates wet scavenging of soluble species

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for convective and stratiform precipitation separately. Nitric acid is used as a reference and the scavenging rate of any other species subject to wet removal is scaled to the scavenging rate of HNO_3 according to the temperature dependent Henry's Law Equilibrium Constant on the basis of Seinfeld and Pandis (1998) using standard condition Henry's Law Coefficients from the literature (Sander, 1999). Table 1 provides a list of all species subject to wet removal in this version of LMDz-INCA. For a more detailed description and evaluation of both the dry deposition and wet scavenging parameterization in LMDz-INCA we refer to the paper of Hauglustaine et al. (2004). In the version of LMDz-INCA used in this study the same schemes have been applied.

3. Model evaluation

We present here a general evaluation of the LMDz-INCA model in its NMHC version using a simulation which is representative of the 1990s. During development the model has been run almost consecutively for more than 20 model years before a spin-up run for this study was initialized. A six months spin-up was then conducted from July to December using the restart files of the last development run. After this spin-up period the model has been run for another 24 consecutive months, the last 12 months have been used in the analysis. This approach was chosen to ensure that even long-lived species such as methane have reached equilibrium.

The large amount of species considered in this model and the still rather sparse observational data available from measurement campaigns only allow us to discuss the major aspects of global tropospheric chemistry in this evaluation of the model performance. Although the spatial distribution and time evolution of more than 80 species are calculated by LMDz-INCA, the discussion shall be focused on the key species and selected NMVOC. CO concentrations are evaluated by comparison with surface climatologies from Novelli et al. (2003), other species simulated by LMDz-INCA including hydrocarbons, acetone, NO_x , PAN, and HNO_3 are compared to observations from various aircraft missions based on the data compilation by Emmons et al. (2000) and ref-

erences therein. Finally, calculated ozone concentrations are evaluated by comparison with climatological ozonesonde observations (Logan, 1999).

3.1. Carbon monoxide

Direct surface emission deriving from fossil fuel combustion as well as biomass burning are the most important CO sources in the atmosphere. In addition, carbon monoxide is produced in situ by oxidation of methane and non-methane hydrocarbons in the entire troposphere.

Figure 1 shows the monthly mean carbon monoxide surface mixing ratio for January and July. CO is most abundant near the surface sources of the tropics and the northern midlatitudes as well as during the northern hemispheric winter when its photochemical lifetime is increased owing to a decrease in the OH abundance and weak vertical mixing. Predicted mixing ratios reach 300 ppb over these regions. During the summer months CO concentrations decrease significantly in the photochemically more active atmosphere, mostly due to a strong increase in OH.

In the tropics the seasonal cycle of CO is controlled to a large extent by the seasonality of biomass burning, which account for half of the direct CO emissions. Biomass burning is most intense during the dry season (December–April in the northern tropics, July–October in the southern tropics). Over these areas with intense biomass burning (tropical regions of Africa and South America) the model calculates maximum mixing ratios in the range of 200 to 300 ppb. In the marine boundary layer of the southern hemisphere, background CO concentrations as predicted by LMDz-INCA are generally lower than 70 ppb.

Simulated monthly mean carbon monoxide concentrations near the surface are compared with climatologies from observations for 18 selected stations from the CMDL network (Novelli et al., 2003) in Fig. 2. Measurements are depicted as monthly means including their standard deviation over the period of record (5 to 10 years, depending on the station).

The model captures well both the absolute values and the seasonal variation of CO

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mixing ratios in most cases, except for a tendency to overestimate CO by up to 20 ppb at southern mid and high latitudes (cf. Cape Grim, South Pole). The generally higher CO concentrations in the northern hemispheric mid- and high latitudes (e.g., Alert, Baltic Sea, Mace Head, Bermuda) are well captured by the model at virtually all sites, both in terms of seasonality and magnitude. The model also shows a good agreement with tropical sites in the northern hemisphere (e.g., Easter Islands, Canary Islands, Midway). The model reproduces the seasonal cycle and spring maximum in the southern tropics (Ascension Island and American Samoa) fairly well, which is governed by the seasonal biomass burning emissions (July to October in the southern hemisphere). However, the peak in CO is one month early at Ascension and overestimated but in phase at the Samoa site.

The comparison suggests that the model is capable of reproducing CO concentration well at locations where well constraint anthropogenic CO sources contribute a significant portion of the emissions (practically all of the northern hemisphere). In the southern hemisphere, where CO concentrations are generally lower and are dominated by biomass burning and, to some extent at southern high latitudes, even by the oceanic source, the model has a tendency of overestimating CO concentrations. This is most likely due to the CO biomass burning emission set, but could also be caused by an overestimate of the VOC emissions with subsequent in-situ formation of CO by photooxidation or an underestimate of the concentration of hydroxyl radicals near the surface at these locations. In general though, the tendency to underestimate CO concentrations, persistent in many current chemistry models both taking and not taking into account VOC photochemistry (cf., e.g., Hauglustaine et al., 1998, 2004; Poisson et al., 2000; Bey et al., 2001), is not discernible in the NMHC version of LMDz-INCA, which we tentatively attribute to the spatial and time variability of the OH abundance well reproduced by the model (cf. Sect. 3.2).

3.2. Hydroxyl radical

OH is most abundant in the tropical lower and mid troposphere reflecting high levels of ultraviolet radiation and water vapor. OH concentrations can reach $20 \text{ to } 30 \times 10^5 \text{ molecules cm}^{-3}$ in this atmospheric domain and decrease with altitude due to a decline in water vapor abundance.

The global mean OH concentration as calculated by the model can be evaluated by using the methylchloroform (CH_3CCl_3) lifetime as a proxy (Spivakovsky et al., 1990; Prinn et al., 1995). Spivakovsky et al. (2000) calculated a global mean atmospheric lifetime of 4.6 years for methylchloroform, in close agreement with Prinn et al. (1995).

Assessing stratospheric and oceanic sinks for CH_3CCl_3 with corresponding lifetimes of 43 and 80 years respectively, Spivakovsky et al. (2000) established a tropospheric methyl chloroform lifetime against OH reaction of 5.5 years. Houweling et al. (1998), Wang et al. (1998a), Mickley et al. (1999), and Bey et al. (2001) obtained corresponding lifetimes of 5.3, 6.2, 7.3, and 5.1 years from their model calculations, respectively. LMDz-INCA calculates a tropospheric CH_3CCl_3 lifetime with respect to oxidation by OH of 5.5 years in excellent agreement with the estimates of Spivakovsky et al. (2000) and Prinn (2001) and well within the range of previously mentioned model studies.

A similar relation persisting between atmospheric methane lifetimes and hydroxyl radical concentrations, OH being the most important atmospheric CH_4 sink, can be used to further evaluate model performance in terms of photochemical activity. The annual mean tropospheric CH_4 lifetime as calculated by the model amounts to 9.2 years. In the 2001 IPCC report (Prather et al., 2001) an estimate for this quantity has been given suggesting a an annual mean tropospheric methane lifetime of 9.6 years. LMDz-INCA is in quite close agreement with this IPCC assessment. Figure 3 summarizes the seasonal cycle of the globally averaged chemical lifetimes of methane and methylchloroform, the hydroxyl radical abundance, and the tropospheric ozone burden. The model calculates global annual mean tropospheric values of $9.6 \times 10^5 \text{ molecules cm}^{-3}$ and 303 Tg for the OH abundance and ozone burden, respectively.

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As already mentioned, the OH abundance to a large extent defines the oxidizing capacity of the atmosphere (cf. Levy, 1971). Using OH concentrations calculated by LMDz-INCA, we give an assessment of the photochemical activity of the troposphere broken down into nine different subdomains based on recommendations by Lawrence et al. (2001). The subdomains are defined as follows: horizontal subdomains are the northern extratropics (NET, 90°–30° N), tropics (TRO, 30° N–30° S), and southern extratropics (SET, 30°–90° S). Vertical subdomains are defined as the planetary boundary layer (PBL, below 750 hPa), free troposphere (FT, 750–500 hPa), and upper troposphere and tropopause region (UT, 500–250 hPa), respectively. The results are summarized in Table 6

From Table 6 it can be seen that the tropical planetary boundary layer has the highest oxidative capacity of all tropospheric subdomains with an annual mean OH abundance of 16.6×10^5 molecules cm^{-3} . The oxidative capacity calculated by the model decreases with increasing altitude, consistent with the current knowledge. Viewed over the entire altitude range, the tropical troposphere shows the highest oxidative capacity, in the southern extratropical troposphere OH concentrations are on average significantly lower.

3.3. Nitrogen compounds

The importance of nitrogen compounds is a consequence of their role in the budget of other key tropospheric species, such as ozone and the hydroxyl radical. Odd nitrogen is mainly released in the form of NO, predominantly as a result of combustion processes (fossil fuel combustion, biomass burning) and soil microbial activity. Global lightning activity also represents a significantly atmospheric source of NO. Once emitted, NO is rapidly converted by photochemical processes into other states of oxidation (NO_2 , NO_3 , N_2O_5) as well as HNO_3 and organic nitrates, including peroxyacetyl nitrate (PAN).

A comparison of observed and calculated profiles of NO_x and PAN are shown in Fig. 4 and 5. The calculated NO_x mixing ratio is generally in good agreement with observations, given the large spatial and time variability in this short-lived species.

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The typically “C-shaped” profiles with higher mixing ratios in the PBL and the upper troposphere as well as decreased mixing ratios in the free troposphere are also well captured. The model, however, overestimates NO_x over Ireland during SONEX below 8 km which could be due to an overestimate of the anthropogenic NO_x source in EDGAR v3.2 or the export of NO_x from the continent being too strong in our model over this region. The model underestimates NO_x concentrations in the upper troposphere during SUCCESS.

Organic nitrates in general, and in particular peroxyacetyl nitrate (PAN), are chemically more stable compounds than NO_x . PAN is the most abundant organic nitrate that has been detected in the atmosphere resulting from a variety of organic precursors, such as isoprene, acetaldehyde, and alkanes (e.g. Roberts, 1990; Altshuller, 1993). Predominantly produced in the PBL by reaction of peroxyacetyl radicals with NO_2 , PAN subsequently becomes subject to long-distance transport to remote environments and to the free troposphere, where it acts as a reservoir of NO_x . Eventually, NO_x is released following thermal decomposition and photolysis (Crutzen, 1979; Kasibhatla, 1993; Moxim et al., 1996). PAN lifetimes vary from a few hours in the lower troposphere, where thermal decomposition effectively limits its residence time, to several months in the free and upper troposphere. LMDz-INCA also considers PAN analogs that derive from higher NMHC as well as other bulk organic nitrate species (cf. Table 1). They generally have shorter atmospheric lifetimes and, hence, significant concentrations are found only in the PBL.

Calculated profiles of the PAN mixing ratio are mostly within one standard deviation of the observations, but show a tendency to underestimate PAN in regions affected by biomass burning as during TRACE-A. A tendency to overestimate PAN over polluted areas as during SONEX is also discernible. The model seems to better reproduce PAN profiles over unpolluted remote regions (PEM-West, PEM-Tropics, TRACE-A) than over regions with a strong continental influence (SONEX, ABLE). Observed PAN concentrations show a maximum between approximately 4 and 8 km depending on the individual location. This mid-tropospheric maximum is a consequence of an increas-

ing lifetime against thermal decomposition with increasing altitude and a decreasing lifetime against photolysis with increasing altitude. The model qualitatively and quantitatively captures this feature well.

Nitric acid (HNO_3) in the atmosphere is produced by reaction of NO_2 with OH and by hydrolysis of N_2O_5 ; major sinks are dry and wet deposition. Comparison of simulated vertical profiles of HNO_3 with observations as shown in Fig. 6 reveals a pronounced tendency of LMDz-INCA to overestimate HNO_3 concentrations, typically by a factor of two. The largest deviations are found in the upper troposphere between 9 and 12 km. This problem is common to most current global 3-D models of tropospheric chemistry (Hauglustaine et al., 1998, 2004; Lawrence et al., 1999; Poisson et al., 2000; Bey et al., 2001). However, in the NMHC implementation of LMDz-INCA this tendency to overestimate HNO_3 concentrations seems to be less pronounced. Bey et al. (2001) have suggested partitioning of nitric acid into aerosols as a possible explanation for this common model deficiency, since most current tropospheric chemistry model do not differentiate between gaseous and aerosol nitrate as is also the case for LMDz-INCA. Studies by Bauer et al. (2004) with a version of LMDz-INCA including such fractionation processes but without NMVOC photochemistry seem to support this argument. A better agreement between calculated HNO_3 concentrations and observations is obtained over remote regions, in particular during the PEM-Tropics campaigns, indicating that HNO_3 overestimation in the model is most pronounced over areas with high NO_x emissions.

3.4. Methanol, acetone, and other VOC

The seasonal mean methanol mixing ratio at the surface is depicted in Fig. 7 for the winter and summer seasons. The principal global CH_3OH source is plant growth. The methanol surface mixing ratio as calculated by the model follow the global seasonal vegetation cycle with low values at mid- and high latitudes in the winter hemisphere, elevated levels in the same latitudinal range in the summer hemisphere, and year-round relatively high mixing ratios over continental areas in the tropics. The methanol

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mixing ratio ranges between 10 and 25 ppbv over tropical South America and Africa during the winter season and reach 30 to 40 ppbv over Southeast Asia and the Eastern United States during the summer season.

Calculated vertical CH₃OH profiles are compared to observations in Fig. 8. The model shows a tendency to underestimate the methanol mixing ratio over remote areas (observations from PEM-Tropics B) where measured values can be higher by up to a factor of 2.5 in the lower troposphere. Generally, the agreement is better at higher altitudes. On the other hand, the model seems to overestimate CH₃OH significantly at northern hemispheric midlatitudes as compared to measurements from the SONEX campaign.

Figure 9 shows a comparison of the seasonal variation in formaldehyde (CH₂O) mixing ratios as calculated by the model with observations near the surface. Observational data has been collected in the framework of the CMDL network. Most of the observational data available has been collected at northern hemispheric midlatitudinal stations, except Zeppelin which is located at polar latitudes. The model shows a fair agreement with the observations at several stations but considerably underestimates formaldehyde concentrations at Ispra, Mace Head, and Zeppelin.

Figure 10 compares simulated formaldehyde profiles to aircraft measurements for various campaigns. Profiles calculated by LMDz-INCA are in good agreement with observations collected during SONEX which are representative of near continental, more polluted regions, as well as with data gathered during the PEM-Tropics B campaign. The latter locations are typical for remote, less polluted areas of the troposphere. The most pronounced deviation between model and measurements occurs for the TRACE-A campaign at the South Atlantic location, where LMDz-INCA seems to overestimate measured formaldehyde concentration in the lower and middle troposphere by up to a factor of three. This disagreement could be related to sources of formaldehyde or its precursors from biomass burning being too high in our emission inventory near this region (South American continental sources).

Vertical profiles of acetone from the model and observations are compared in Fig. 11.

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Over most regions (including those sampled during PEM-West B and PEM-Tropics B) the model estimates of the acetone mixing ratio agree fairly well with the observations. The higher levels observed during the TRACE-A campaign over Brazil and Africa, also well simulated by the model, are related to acetone emissions from biomass burning.

5 The model captures the vertical gradients in most cases. It generally also reproduces the acetone mixing ratios observed in the Asian outflow during PEM-WEST B, but significantly overestimates acetone at the near continental sites measured during the SONEX campaign. However, acetone profiles calculated by LMDz-INCA do agree quite well with observations obtained during the ABLE-3B (Singh et al., 1994a) experiments.

10 The observations gathered during PEM-Tropics B indicate surprisingly large abundances of acetone over the tropical Pacific (Singh et al., 2001), a region where most models previously predicted quite low concentrations (cf., e.g. Hauglustaine et al., 1998; Bey et al., 2001; Horowitz et al., 2003). These observations suggest the presence of a large natural, distributed source of oxygenated organic species as has been discussed by Horowitz et al. (2003). The current version of LMDz-INCA does not seem to show the same disagreement, possibly owing to the inclusion of an oceanic source for acetone based on the Jacob et al. (2002) recommendations.

3.5. Ozone

Figure 12 shows seasonal mean ozone mixing ratios calculated near the surface for the winter and summer seasons. Maximum mixing ratios reaching 50 to 70 ppbv are calculated over industrialized regions of the northern hemisphere during summertime, where strong biogenic NMVOC sources combine with pronounced NO_x concentrations (Eastern United States, Europe, Southeast Asia). Elevated ozone levels of up to 50 ppbv are also visible over areas with intense biomass burning emissions in the tropical continental domain. In general, the ozone distribution as calculated by LMDz-INCA is in good agreement with previous studies (cf. e.g. Hauglustaine et al., 1998; Wang et al., 1998b; Poisson et al., 2000; Bey et al., 2001; Horowitz et al., 2003).

Over the oceans the ozone background mixing ratio is generally less than 25 ppb

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in both hemispheres except over the North Pacific and the North Atlantic during wintertime, where export of O_3 from the continental boundary layer is clearly visible. In regions remote from any pollution sources, such as the Central Pacific Ocean, the Indian Ocean, or the Southern Ocean, ozone mixing ratios can drop to less than 15 ppb. A summertime minimum with less than 15 ppbv of ozone is calculated over the North Pacific, where NO_x concentrations are very low while ozone destruction by photolysis is significant.

Figure 13 provides an evaluation of the surface ozone mixing ratio as derived by the model by comparison with measurements at 15 selected stations. Observational data have been taken from the CMDL network (cf. Komhyr et al., 1989; Oltmans et al., 1989; Oltmans and Levy II, 1992, 1994) and represent monthly means, averaged over the total observational record at each station, which cover between 5 and 10 years. In general, calculated monthly mean ozone mixing ratios as well as the simulated seasonal cycle are in good agreement with the observed values at most sites. The simulated ozone mixing ratios reproduce generally well the observed values in subtropical regions and at midlatitudes in the Northern Hemisphere (Mace Head, Payerne, Hohenpeissenberg, Sable Island, Zugspitze, Mauna Loa). The model fails to capture the arctic springtime ozone minimum at Barrow, which has been attributed to bromine catalyzed ozone depletion events (Barrie et al., 1988; Tang and McConnell, 1996; Martinez et al., 1999; Wennberg, 1999). In the Southern Hemisphere, the seasonal cycle and magnitude of ozone is fairly well reproduced at subtropical- (American Samoa) and mid latitudes (Cape Grim), but both the seasonality and magnitude are considerably underestimated at South Pole during most of the year (April to November). As photochemistry is slow in the polar regions, a connection of this discrepancy with transport processes seems likely.

The vertical distribution of the ozone mixing ratio as calculated by LMDz-INCA is evaluated against ozone soundings in Fig. 14. The observations are taken from the climatology described by Logan (1999). The comparison confirms the general good agreement with the observations at most stations. With only a few exceptions, model

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results are within 10 to 15 ppb of the observed values, they are well within their observed variability ($\pm 1\sigma$ standard variation), and show the correct seasonal phase.

At high northern latitudes (Alert station) the model captures the spring maximum in the lower troposphere and the maximum from May to August in the middle troposphere.

The model underestimates, though, the monthly mean mixing ratios during spring and summer near the tropopause and also visibly underestimates the observed variability. The weakly developed variability in simulated upper tropospheric ozone as compared to the observations has been attributed to an overestimate of the tropopause height by the GCM (Hauglustaine et al., 2004) resulting in lower than observed O_3 mixing ratios in the above 300 hPa region.

Seasonal variations at northern mid-latitudes are fairly well reproduced (Hohenpeisenberg, Payerne). The timing of the ozone maximum in polluted regions (spring in the upper troposphere and a gradual shift toward the summer season with decreasing altitude) is also generally well reproduced by the model to within the constraints offered by the observations at mid-latitude stations (Sapporo, Payerne, Boulder). A good model-to-measurement agreement over the entire troposphere is achieved at tropical (American Samoa) and at southern midlatitude (Asperon) stations.

Table 7 provides an estimate of the annual ozone budget in the troposphere (defined as the domain extending from the surface up to 200 hPa) as calculated with LMDz-INCA. The results indicate that sources and sinks in the troposphere are dominated by photochemistry. The global photochemical ozone production and destruction amount to 4436 and 3890 $Tg O_3 yr^{-1}$, respectively. Clearly, gross photochemical production by far exceeds stratospheric influx, which contributes 715 $Tg O_3 yr^{-1}$ to the total tropospheric budget. Net tropospheric photochemical production of ozone, calculated as the difference between gross production and destruction, totals 546 $Tg O_3 yr^{-1}$ which is slightly more than 75% of the stratospheric influx.

According to the calculations with LMDz-INCA, 1261 $Tg O_3 yr^{-1}$ are removed from the troposphere by dry deposition at the surface. Dry deposition of ozone is almost two times more effective in the Northern Hemisphere (835 $Tg O_3 yr^{-1}$) than in the South-

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ern Hemisphere ($426 \text{ Tg O}_3 \text{ yr}^{-1}$). This can be explained by the fact that the northern hemispheric continental area substantially exceeds the southern hemispheric land mass and by the rather weak solubility of ozone in water which significantly limits ozone surface deposition rates over the oceans. The average photochemical ozone lifetime calculated on the basis of the tropospheric ozone burden of 303 Tg and the photochemical destruction is 28.4 days. Taking into account physical removal of ozone by dry deposition, the tropospheric O_3 lifetime is 21.5 days on global average.

Photochemical production of ozone is much more pronounced in the Northern Hemisphere ($2773 \text{ Tg O}_3 \text{ yr}^{-1}$) than in the Southern Hemisphere ($1663 \text{ Tg O}_3 \text{ yr}^{-1}$), likely due to the significant bias in NO_x sources toward the Northern Hemisphere related to a higher degree of industrialization as well as the overall stronger biogenic NMVOC surface flux from the terrestrial vegetation, the latter a consequence of the larger land mass in the Northern Hemisphere. More than 60% of the gross production occurs in the Northern Hemisphere and about than 85% of the net photochemical ozone production arises within the same hemisphere.

Previous studies with global 3-D models indicate photochemical production rates ranging from 3314 to $4550 \text{ Tg O}_3 \text{ yr}^{-1}$ and photochemical loss rates in the range between 2511 and $4065 \text{ Tg O}_3 \text{ yr}^{-1}$ (World Meteorological Organisation, 1998; Lelieveld and Dentener, 2000; Prather et al., 2001). Given these numbers, net photochemical production is estimated to amount between 485 and 803 Tg yr^{-1} . Stratospheric inputs in these models vary from 390 to $768 \text{ Tg O}_3 \text{ yr}^{-1}$ (World Meteorological Organisation, 1998). The estimates calculated by LMDz-INCA are well within the range given by previous models, even though they reside on the high end. However, they are fairly close to previous results obtained with the CH_4 -only version of LMDz-INCA (Hauglustaine et al., 2004), with the exception of stratospheric influx ($715 \text{ Tg O}_3 \text{ yr}^{-1}$ as opposed to $523 \text{ Tg O}_3 \text{ yr}^{-1}$ in the LMDz-INCA- CH_4 version) and a substantially more pronounced imbalance in the net photochemical production of ozone between the two hemispheres (469 and $77 \text{ Tg O}_3 \text{ yr}^{-1}$ for the northern and southern hemisphere, respectively, as compared to 451 and $116 \text{ Tg O}_3 \text{ yr}^{-1}$ in the LMDz-INCA- CH_4 version). The model cal-

culates a global mean O₃ burden of 303 Tg O₃ which divides into a hemispheric burden of 178 Tg O₃ (59%) and 125 Tg O₃ (41%) for the northern and southern hemisphere, respectively.

4. Results and discussion

The results presented in this section are based on a control run which is identical to the run used in the model evaluation. As mentioned above, this control run has been taken over 30 model months including a six months spin-up. The last 12 months have been used in the analysis. In addition, we conducted several model experiments to study various aspects of the impact of NMVOC on tropospheric chemical composition and photochemistry. These experiments are focused on the impact of biogenic NMVOC which have their predominant or unique source in the terrestrial biosphere, such as for instance isoprene or methanol.

To quantify the impact of these compounds, four experiments have been carried out assessing the impact of isoprene, biogenic methanol and acetone, as well as terpenes. In these experiments the respective emissions were turned off separately and changes in key tropospheric species, such as ozone, OH, NO_x and PAN, and HO_x have been analyzed. We note that non-linearities in photochemistry can somewhat bias the results of these sensitivity experiments affecting the oxidation capacity of the atmosphere, likely towards an increase in the reactivity (measured by the OH concentration) when turning off the emissions, because the highly reactive biogenic VOC are a strong OH sink. To quantify the uncertainties associated with the above mentioned non-linearities, we have performed an additional experiment with isoprene emissions reduced to 90% of their standard magnitude and compared the results to the “standard” experiment with zero isoprene emissions. All experiments were initialized with the July restart files of the first year of the control run and then continued over 18 model months. The last 12 months were used for comparison.

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4.1. Impact of isoprene

The major impact of isoprene on tropospheric photochemistry and chemical composition lies in its role as an ozone precursor and in the formation of organic nitrates in the course of isoprene photooxidation. Poisson et al. (2000) have shown that organic nitrate formation in the course of NMVOC oxidation is responsible only for a minor decrease of a few percent in NO_x concentrations over polluted regions. The impact of NMVOC oxidation via the formation of PAN and analogues significantly increases NO_x concentrations in remote environments with only feeble NO_x emissions, such as the marine PBL. Thermal decomposition of PAN can increase NO_x concentrations by almost a factor of 2, which is enough to significantly impact ozone and OH formation in these remote areas.

Figure 15 summarizes the impact of isoprene on ozone and carbon monoxide mixing ratios at the surface for January and July. During the northern hemispheric winter changes in the ozone mixing ratio near the surface due to the presence of isoprene are between 1 and 2 ppbv globally. Calculations with LMDz-INCA show a maximum increase of ozone mixing ratios in January of up to 10 ppbv over eastern equatorial Africa and some regions of South America, but in general the impact of isoprene on ozone production during the northern hemispheric winter seems to be small and globally homogenous. In contrast, the model predicts important changes in the ozone concentration over the Eastern United States, Europe, and, to some extent, over Southeast Asia for July. Over the eastern US adding isoprene emissions results in an increase in ozone at the surface by up to 30 ppbv. The enhancement in ozone burden in July seems to be a direct consequence of high levels of NO_x prevailing in this region and the strong surface source of isoprene.

The changes in the ozone burden of the troposphere due to the biogenic isoprene source potentially could affect the radiative budget. To quantify the magnitude of the ozone net radiative forcing as a result of increased ozone levels we performed an offline radiative transfer calculation using the ozone fields provided by the control run and the

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isoprene experiment. This calculation predicts a global mean net radiative forcing of ozone of 0.09 W m^{-2} . The ozone net radiative forcing is most pronounced at tropical latitudes ranging between 0.07 and 0.15 W m^{-2} . Maximum values of up to 0.17 W m^{-2} are found over the desert regions of northern Africa and the Arabic Peninsula. At mid- and high latitudes of both hemispheres the O_3 net radiative forcing ranges between 0.02 and 0.09 W m^{-2} .

Apparently, isoprene photooxidation is a significant source of carbon monoxide. As shown in the lower two panels of Fig. 15, isoprene oxidation near the surface results in an increase of carbon monoxide mixing ratios ranging from 25 to 60 ppbv over the major isoprene source areas, such as the Eastern United States, equatorial Africa, tropical South America, and, to a lesser extent, Southeast Asia, for both, January and July. Atmospheric transport of CO away from these source areas has a significant impact on the equatorial regions of the marine planetary boundary layer. A substantial CO entrainment to remote and relatively unpolluted regions by the trade winds can be seen over the equatorial Atlantic and Pacific ocean enhancing CO levels in the marine PBL between 15 and 35 ppbv. A marked increase in CO mixing ratios ranging from 30 to 65 ppbv is calculated by the model for the Eastern US and Europe during summer.

The changes in surface NO_x mixing ratios due to the presence of isoprene are depicted in the upper two panels of Fig. 17 for January and July. The impact of isoprene emissions is apparent on the global scale. Calculations with LMDz-INCA show a strong decrease in NO_x mixing ratios over regions with generally high levels of NO_x (Eastern United States and Europe) for both January and July. A decrease in NO_x mixing ratios between 50 and 250 pptv in January is discernible over these regions. In July NO_x levels can even be reduced by up to 1000 pptv over the same regions due to the strong isoprene sources during the summer months.

Once NO_x has been converted into PAN and analogue compounds which possess much longer lifetimes than nitrogen oxides themselves, NO_x is redistributed globally. Figure 17 shows an increase in the NO_x mixing ratio of up to 100 pptv in January over the tropical oceans in the vicinity of the western continental regions. In July LMDz-INCA

calculations seem to indicate that NO_x is preferably released closer to the regions with high levels of nitrogen oxides, in particular over the northern hemispheric continental areas. This seems to be a consequence of the the shorter PAN lifetimes during the northern hemispheric summer months.

5 The lower two panels in Fig. 17 increases in PAN mixing ratios at the surface due to isoprene for January and July. During the northern hemispheric winter a homogeneous increase in PAN over the entire extratropical northern hemisphere ranging from 10 to 50 pptv is calculated by LMDz-INCA. This ubiquitous enhancement seems to be a consequence of the longer PAN lifetimes during the winter when thermal decomposition
10 is less effective. The increase in PAN in January is particularly strong close to the primary isoprene sources of equatorial Africa, tropical South America, and Southeast Asia. Over these regions the presence of isoprene enhances PAN mixing ratios by up to 250 pptv.

In July the region of strong enhancements in the PAN surface mixing ratios is shifted
15 to the northern hemispheric midlatitudes. An increase in PAN between 75 and 750 pptv is discernible over most of the United States, Europe, and extratropical Southeast Asia. Over those regions, distinctive isoprene emissions combine with high levels of NO_x from technological sources. This pronounced increase in PAN mixing ratios is even more remarkable since PAN lifetimes are significantly shorter during the northern hemi-
20 spheric summer months. A homogeneous enhancement in PAN mixing ratios, albeit somewhat less pronounced, is apparent over southern hemispheric mid- to high latitudes in July, similar to the PAN increase during northern hemispheric winter months. The development of this extratropical “PAN plume” during the southern hemispheric winter owes its existence to the same extension in PAN lifetimes but it is less distinctive
25 due to significantly lower NO_x sources in the Southern Hemisphere. The increase in PAN mixing ratios in the tropics is less pronounced in July as well reaching a surplus of up to 100 pptv over the tropical regions of South America and Africa.

The major sink of non-methane VOC in the troposphere is the reaction with hydroxyl radicals. As a result it is expected that NMVOC have a significant impact on the global

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distribution of OH and, hence, the tropospheric oxidizing capacity. Isoprene, and all other NMVOC that are emitted in significant amounts, will compete for the OH radical. Figure 16 shows the impact of global biogenic isoprene emissions on OH. LMDz-INCA calculates a reduction in the OH concentration reaching 50% in the planetary boundary layer over the tropics and northern hemispheric mid-latitudes where isoprene emissions are strongest. A decrease in OH ranging between 5% and 25% is also apparent in the free troposphere over the tropical and northern hemispheric mid- and high latitudes. This reduction in OH is most likely caused by isoprene photooxidation products such as methacrolein and formaldehyde which possess a substantially longer photochemical lifetime. The model also calculates an isoprene induced reduction of the global mean tropospheric OH concentration by approximately 0.9×10^5 molecules cm^{-3} or roughly 10%.

Apparently, isoprene emissions result in a slight increase in OH concentrations between 2.5% and 10% over the entire troposphere at southern mid- to high latitudes. This increase in OH away from the major isoprene sources could be due to the production of PAN which would yield in a redistribution of NO_x to these remote regions of the troposphere. As a result, an increase in the ozone and, hence, also OH production is expected.

Another important consequence of a reduced oxidative capacity of the troposphere is the change in the lifetime of methane. The CH_4 tropospheric lifetime is predominantly determined by its most important sink, the reaction with hydroxyl radicals. Calculations with LMDz-INCA indicate an increase in the global annual mean methane lifetime averaged over the entire tropospheric domain by approximately 4 months as a consequence of global biogenic isoprene emissions.

The major products of isoprene photooxidation are methy vinyl ketone (MVK), methacrolein (MACR), and formaldehyde (cf. e.g. Sprengnether et al., 2002, and references therein). As discussed in detail by Apel et al. (2002) the ratio of the oxidation products MVK and MACR holds information about the history of the isoprene oxidation. Production of MVK is favored by the reaction of isoprene with the OH radical

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whereas reaction of isoprene with ozone and the nitrate radical (NO_3), during night-time in high NO_x environments, favors MACR as the main primary product. A typical daytime MVK/MACR ratio of approximately 1.4 is expected from reaction kinetic considerations (Apel et al., 2002) for environments in which NO_x is relatively abundant and isoprene is predominantly oxidized by the hydroxyl radical. In addition, MACR reacts more rapidly with OH resulting in a shorter MACR lifetime and higher MVK/MACR ratios under the same photochemical conditions. In NO_x limited environments the predominant products of the isoprene-OH reaction pathway are organic peroxides (Miyoshi et al., 1994) with increased methacrolein yields due to peroxide recycling and consequently MVK/MACR ratios lower than 1.4. Due to the short photochemical lifetime of isoprene of only a few hours the MVK/MACR ratio reflects the aging of the isoprene-MVK-MACR mixture yielding potentially higher MVK/MACR ratios at some distance away from the isoprene sources.

Figure 18 depicts the annual mean MVK/MACR ratio at the surface as calculated by LMDz-INCA. Apparently, large portions of the globe are characterized by MVK/MACR ratios smaller than the 1.4 threshold value. The lower than 1.4 MVK/MACR ratios over the main continental isoprene source regions in tropical South America, Africa, and South-East Asia indicate that isoprene oxidation by ozone and, to a lesser extent, by NO_3 is at least equally important for the isoprene oxidation.

High ratios over the tropical ocean close to the continents reaching values of up to 3 are likely due to photochemical aging of air since isoprene concentrations are negligible and ozone and NO_3 concentrations are low favoring MVK over MACR. MVK/MACR ratios ranging between 1.0 and 1.2 over the open oceans at tropical latitudes seem to be a consequence of prevailing low NO_x concentrations resulting in low levels of ozone and NO_3 as well as the small but non-negligible oceanic isoprene source taken into account by our emission inventory (cf. Table 1). Note that LMDz-INCA reproduces MVK/MACR ratios higher than 1 over the continental isoprene source regions (e.g., the United States, or Europe) which provides an additional check of the chemical mechanism, since current chemistry models often fail to reproduce these elevated

MVK/MACR ratios.

At the extratropical latitudes of both hemisphere MVK/MACR ratios drop below 1.0. These prevailing low MVK/MACR ratios are the result of relatively low concentrations of the hydroxyl radical as compared to tropical latitudes and a moderate ozone abundance. Methyl vinyl ketone exhibits a higher reaction rate with ozone than methacrolein (Atkinson et al., 1997; Sander et al., 2002) which favors MACR over MVK resulting in low MVK/MACR ratios. We note, however, that isoprene, methacrolein, and methyl vinyl ketone concentrations are generally low at these latitudes, at least on the global scale. Isoprene surface emissions are minor, with the exception of the eastern United States during the warm season, which also limits the abundance of the primary oxidation products MVK and MACR.

4.2. Impact of acetone

Acetone is photochemically produced by propane oxidation. Further important precursors are iso-alkanes, iso-alkenes, and terpenes. Moreover, acetone has surface emissions from anthropogenic sources, biomass burning, and the terrestrial vegetation (Warneke et al., 1999). Jacob et al. (2002) have presented a global acetone budget claiming the ocean as an additional important primary source. Acetone is destroyed by reaction with OH and photolysis. Lifetimes range from about 10 days to a few months varying with location and season (Singh et al., 1994b). Photolysis of acetone has been reported a significant source of HO_x radicals in the upper troposphere (Jaeglé et al., 2001).

Figure 19 shows the impact of global acetone surface emissions from natural sources (terrestrial vegetation and the ocean, 47 TgC yr⁻¹) on concentrations of O₃, CH₂O, OH, and HO_x (HO_x=OH+HO₂+∑RO₂). Changes in the ozone concentration are limited to the upper troposphere/lower stratosphere (UT/LS) region and are generally small. LMDz-INCA calculates an increase of O₃ of generally less than 1 ppbv. On the other hand, acetone surface emissions seem to have a significant impact on the concentration of OH and HO_x in the UT/LS region. Figure 19 shows and increase in OH of up to

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10% in the tropical and midlatitudinal UT. Positive OH anomalies due to acetone exceed 20% in the Antarctic UT/LS. The changes in HO_x are even more distinct. An increase of up to 15% prevails in the entire UT/LS region exceeding 20% at polar latitudes. The acetone sensitivity experiment shows basically no impact on ozone, OH, or HO_x in the lower troposphere. Acetone surface emissions also seem to have an almost negligible impact on global formaldehyde concentrations. The model calculates changes in CH_2O concentrations of generally less than 10 pptv for the entire troposphere. The maximum impact is found for the lower troposphere.

4.3. Impact of methanol

Due to its relatively long lifetime, a large amount of methanol can be transported from the PBL into the free troposphere and has a potentially significant impact on the concentration of oxidants. We conducted a sensitivity experiment designed to study this impact of global methanol surface emissions from natural sources (terrestrial vegetation and plant decay, 100 TgC yr^{-1}) on ozone levels, formaldehyde concentrations, and OH as well as HO_x mixing ratios. The results of this experiment are summarized in Fig. 20.

An increase in the ozone concentration is visible over the entire free troposphere, albeit relatively small, and shows two maxima at about 30 degrees latitude in both hemispheres. A maximum increase of 2 ppbv is calculated at the northern hemispheric subtropical tropopause. concentrations. The calculations with LMDz-INCA also show an increase in OH on the order of one to two percent only. On the other hand, these calculations predict enhancements in the level of peroxy radicals, such as HO_2 and organic peroxy radicals, combined to HO_x , between 8 and 15% in the UT/LS region depending on latitude. The increased supply of peroxy radicals seems to drive the changes in ozone mixing ratios. The large increase in HO_x in the UT/LS region could also be a consequence of CH_2O photolysis. LMDz-INCA calculates a pronounced increase of formaldehyde in the lower troposphere between 50° S and 50° N up to about 700 hPa with maximum positive anomalies of 70 pptv in the PBL at about 40° N . At these al-

titudes formaldehyde photolysis is relatively weak and consequently CH₂O lifetimes will be longer. The impact of methanol emissions on formaldehyde levels decreases with increasing altitude. The highly efficient destruction of formaldehyde by photolysis seems to be the reason for the very small impact of methanol in the upper troposphere.

5 4.4. Global carbon monoxide budget for the troposphere

LMDz-INCA includes a set of tagged model tracers pertaining to the conversion of primary emitted VOC and intermediate products of tropospheric photooxidation to carbon monoxide and CO₂ (Folberth et al., 2005). Using these tagged tracers we present an updated estimate of the global CO budget and compare our results to the estimates given by Bergamaschi et al. (2000). Table 8 summarizes both studies. Note that our estimates are based on the assumption that global sources and sinks are in perfect equilibrium. In particular, this pertains to the tropospheric-stratospheric exchange of CO which has been derived as the residual of all other sources and sinks calculated directly with the model. It should be pointed out as well that we have not attempted to calculate uncertainty limits for our CO budget. Uncertainties in the global CO source are expected to be high and will depend primarily on the uncertainties in the global emission inventory. The uncertainties in the global sink are a consequence of the source uncertainties and are also expected to be substantial.

Whereas the range of CO sources and sinks for individual tropospheric processes in the study by Bergamaschi et al. (2000) represent a posteriori estimates obtained by an inverse modelling approach, the current CO budget has been derived with a forward model. Furthermore, the studies are based on fundamentally different emission inventories. Nevertheless, both modelling studies agree surprisingly well in their predictions of the total global CO sources and sinks. Note that in the LMDz-INCA standard emission inventory (cf. Table 5) CO emissions deriving from agricultural waste burning and domestic biofuel use have been attributed to the global biomass burning source. In the study of Bergamaschi et al. (2000) these sources are included in the technological source category. For the purpose of this comparison we limited the CO biomass burn-

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ing source to savannah and forest fires only attributing carbon monoxide emissions from agricultural waste burning and domestic biofuel use to the technological sources. With these changes in categories, the carbon monoxide budget calculated by LMDz-INCA and summarized in Table 8 appears to be in very good agreement with the study of Bergamaschi et al. (2000).

Secondary CO sources in LMDz-INCA deriving from reactive carbon compound (RCC) photooxidation in the troposphere amount to $1489 \text{ TgCO yr}^{-1}$. RCC include all carbon containing compounds that are chemically broken down in the troposphere, excluding only CO_2 which is chemically inert. Oxidation of methane, isoprene, methanol, terpenes, acetone, and other non-methane VOC (mostly of anthropogenic origin), respectively, contribute 854, 359, 110, 49, 28, and 89 TgCO yr^{-1} to this secondary carbon monoxide source. In general, conversion efficiencies – defined as the ratio of the amount of carbon emitted in the form of individual species to the amount of carbon converted to CO expressed in percent – are less than 50% (20%–45%) for all species other than methane and CO. Therefore, it follows that these species must possess other highly efficient sinks. These sinks are found to be the physical removal of primary emitted VOC as well as their stable intermediate photochemical products (such as alcohols, aldehydes, ketones, and organic nitrates) through surface dry deposition and wet removal. The CH_4 conversion efficiency of 90% indicates that photochemical destruction is by far the most important atmospheric sink, CH_4 entrainment to the stratosphere being the only other significant removal process.

Both studies agree closely on the global CO sink. Photochemical oxidation of carbon monoxide, which represents a significant secondary CO_2 source (Folberth et al., 2005), constitutes a slightly stronger sink in LMDz-INCA. Calculations with LMDz-INCA indicate a global CO-to- CO_2 conversion efficiency of 88%. On the other hand, dry deposition of CO at the surface seems to be significantly weaker in LMDz-INCA than in the study of Bergamaschi et al. (2000). The residual of all sources and sinks of carbon monoxide in our model is interpreted as a CO entrainment to the stratosphere. The estimate of this particular sink significantly exceeds the estimate of Bergamaschi et al.

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(2000) as a consequence of the substantially weaker CO surface deposition.

5. Summary and conclusions

In this paper we provide a first description and evaluation of a new version of the 3-D-chemistry-climate model LMDz-INCA. This new version is based on the recently developed CH₄-NO_x-CO-O₃ version of LMDz-INCA to which a comprehensive representation of the photochemistry of non-methane hydrocarbons (NMHC) and volatile organic compounds (VOC) has been added. The current version runs with a horizontal resolution of 3.8×2.5 degrees in longitude and latitude, respectively, and the vertical model domain extends over 19 levels from the surface up to 3 hPa. The model calculates the distribution of tropospheric ozone and its precursors as well as a substantial number of primary NMVOC and intermediate photochemical products, such as isoprene, acetone, methanol, PAN and analogues, or formaldehyde. Surface emissions are based on state-of-the-art emission inventories. Biogenic VOC emissions from the terrestrial vegetation are calculated with the dynamic vegetation and emission model ORCHIDEE and are provided off-line to LMDz-INCA. The model is run in climatological mode and the results are evaluated by thorough comparison with observations inasmuch as they are available.

The model captures well the distribution and seasonal cycle of carbon monoxide except for a tendency to overestimate CO by up to 20 ppb at southern mid and high latitudes. The tropospheric annual mean methane (9.2 years) and methylchloroform (5.5 years) chemical lifetimes are in excellent agreement with estimates obtained by observations indicating that LMDz-INCA simulates the global tropospheric OH distribution and magnitude (annual mean OH abundance of 9.6×10^5 molecules cm⁻³) fairly well. A comparison with observed vertical profiles of NO_x, PAN, and HNO₃ indicate that LMDz-INCA is capable of well reproducing the distribution of nitrogen compounds, even though the problem of overestimated nitric acid concentrations, prevailing in many current chemistry models, also persists in LMDz-INCA, albeit to a somewhat reduced

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extent.

The model also satisfactorily simulates a variety of non-methane volatile organic compounds, such as acetone, methanol, and formaldehyde. The comparison with observations reveal a generally good agreement for these species. Locally, however, methanol can be underestimated by up to a factor of 2.5 at some remote locations. The comparison also indicates that the model overestimates measured formaldehyde concentrations in the lower and middle troposphere over the South Atlantic by up to a factor of three. This disagreement could be related to an overestimate of formaldehyde or its precursors from biomass burning being too high in our emission inventory near this region (South American continental sources). The comparison of acetone profiles with observations show a fairly well model-to-data agreement over the entire troposphere.

The model reproduces fairly well the magnitude and distribution of ozone in the entire troposphere. Annual mean tropospheric photochemical production ($4436 \text{ Tg O}_3 \text{ yr}^{-1}$), and destruction ($3890 \text{ Tg O}_3 \text{ yr}^{-1}$), dry deposition at the surface ($1261 \text{ Tg O}_3 \text{ yr}^{-1}$), and stratospheric influx ($715 \text{ Tg O}_3 \text{ yr}^{-1}$) are in good agreement with analogue estimates of current photochemical models. A net tropospheric ozone production rate of $546 \text{ Tg O}_3 \text{ yr}^{-1}$ is calculated by LMDz-INCA which is about 75% of the the stratospheric influx.

The new version of LMDz-INCA has been applied to study the impact of isoprene (411 TgC yr^{-1}), acetone (50 TgC yr^{-1}), and methanol (104 TgC yr^{-1}) surface emissions on ozone and other tropospheric oxidants. Isoprene emissions are found to have a substantial impact on ozone and carbon monoxide concentrations. The calculations with LMDz-INCA show a strong increase of up to 30 ppbv at the surface over the Eastern United States and Europe in the summer as a consequence of the interaction of high levels of both isoprene and NO_x . To assess the potential importance of these changes in the tropospheric ozone burden on the radiative budget of the atmosphere, an offline radiative calculation has been conducted. A global mean ozone net radiative forcing of 0.09 W m^{-2} is calculated with maximum values reaching 0.17 W m^{-2}

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at tropical latitudes. The radiative effect of an increasing ozone burden in the troposphere as a result of biogenic isoprene surface emissions on the atmospheric radiative budget, therefore, appears to be non-negligible but rather small as compared to other seemingly more important effects. However, an increase in the source of tropospheric ozone precursors, such as NO_x and BVOC originating from a more active biosphere, as a consequence of a changing climate could potentially represent a positive feedback process enhancing global warming. Clearly, a more detailed investigation is needed.

Carbon monoxide mixing ratios are found to increase between 25 and 60 ppbv over continental isoprene source regions and between 15 and 30 ppbv in the remote marine boundary layer. The key role of isoprene lies in its impact on the redistribution of nitrogen oxides in the entire troposphere due to the conversion of short-lived NO_x into organic nitrates with significantly longer lifetimes. Photooxidation of isoprene provides large amounts of organic nitrate precursors. An increase in PAN mixing ratios at the surface ranging from 10 to 250 pptv and 75 to 750 pptv during northern hemispheric winter and summer months, respectively, is calculated by LMDz-INCA. A decrease in the zonally averaged annual mean OH concentration ranging between 5% and 25% in the tropical and northern hemispheric free troposphere is calculated by our model as a consequence of the global biogenic isoprene source. An isoprene induced reduction of the tropospheric OH concentration by approximately 0.9×10^5 molecules cm⁻³ or roughly 10% is calculated on the global average. Ensuing this reduction in the oxidative capacity the model predicts an increase of the global mean tropospheric methane lifetime by approximately four months. Our simulations reveal MVK/MACR ratios of less than 1.4 expected from reaction kinetic considerations for large portions of the globe suggesting that reaction of isoprene with ozone and NO₃ is equally important for the oxidation of isoprene in the lower troposphere.

Acetone and methanol play a significant role in controlling tropospheric oxidants in the upper troposphere/lower stratosphere region. The sensitivity experiments show a small impact of global acetone surface emissions on tropospheric ozone but indicate a strong impact on the OH and peroxy radical budget of the UT/LS region. An increase

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between 10% and 15% in the UT/LS in the OH and HO_x concentration is calculated by the model. Interestingly, acetone seems to have the most pronounced impact over the polar regions, at least when measured as percent changes. In case of the methanol sensitivity experiments, calculations show maximum positive anomalies of 2 ppbv for ozone, 2% for OH, and 8 to 15% for HO_x, respectively. Both species show relatively small impacts on tropospheric formaldehyde concentrations with maximum changes of 10 pptv and 70 pptv for acetone and methanol, respectively. Both these maxima occur in the lower troposphere or even in the PBL.

We also attempted an estimate of the CO budget for the global troposphere and compared the results to a recently published inverse modelling study. Even though our study and the study by Bergamaschi et al. (2000) have used fundamentally different approaches to derive their estimates of global tropospheric CO sources and sinks (a forward calculation in LMDz-INCA versus a posteriori estimates obtained by means of an inverse modelling approach in the study Bergamaschi et al., 2000) both studies show a strikingly similar global picture. A global carbon monoxide source of 3019 TgCO yr⁻¹, which divides into a primary CO source of 1533 TgCO yr⁻¹ originating from biomass burning, technological, and oceanic sources and a secondary CO source of 1489 TgCO yr⁻¹ due to methane and VOC photooxidation, is calculated by LMDz-INCA. Oxidation of methane, isoprene, methanol, terpenes, acetone, and other non-methane VOC, respectively, are found to contribute 854, 359, 110, 49, 28, and 89 TgCO yr⁻¹ to this secondary CO source. The model calculates global VOC-to-CO conversion efficiencies ranging from 20 to 45%, a 90% conversion efficiency is prognosticated in case of methane. Oxidation of CO represents by far the most important tropospheric sink. According to our calculations, 2653 TgCO yr⁻¹ are removed from the troposphere by this process indicating a global CO-to-CO₂ conversion efficiency of 88%. Surface deposition of CO (135 TgCO yr⁻¹) and stratospheric CO entrainment (231 TgCO yr⁻¹) are identified as the only significant tropospheric carbon monoxide sinks other than CO oxidation.

In summary, the NMHC version of LMDz-INCA provides an overall satisfactory rep-

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resentation of the complex photochemical processes of the tropospheric VOC composition and. The new version has been applied in several sensitivity studies addressing the interaction between the continental biosphere and atmospheric chemical composition. Current development efforts are focused on halogen photochemistry, an interactive climate-chemistry coupling, and integration of LMDz-INCA with the dynamic vegetation and emission model ORCHIDEE providing a comprehensive modelling tool to study more directly the numerous feedbacks between the biosphere, the atmosphere, climate, and photochemistry. Results from these new versions will be reported in forthcoming studies

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Table 1. Chemical species in LMDz-INCA.

#	tracer	family ^a	description	ch. ^b	a/c ^c	em. ^d	d.d. ^e	w.s. ^f	τ ^g
1	O _x	O ₃ +O(¹ D)+O(³ P)	odd oxygen (ozone and atomic oxygen)	•	•	–	•	–	s
2	O ₃ l	single	inert ozone ^h	–	•	–	•	–	l
3	O ₃ S	single	stratospheric ozone ⁱ	–	•	–	•	–	l
4	N	single	nitrogen radical	•	–	–	–	–	s
5	N ₂ O	single	nitrous oxide	•	•	•	–	–	l
6	NO	single	nitric oxide	•	•	• ^j	•	–	s
7	NO ₂	single	nitrogen dioxide	•	•	–	•	–	s
8	NO ₃	single	nitrate radical	•	•	–	•	–	s
9	N ₂ O ₅	single	nitrogen pentoxide	•	•	–	•	–	s
10	HNO ₂	single	nitrous acid	•	•	–	•	•	s
11	HNO ₃	single	nitric acid	•	•	–	•	•	s
12	HNO ₄	single	peroxynitric acid	•	•	–	•	•	s
13	H	single	hydrogen radical	•	–	–	–	–	s
14	H ₂	single	molecular hydrogen	•	•	•	•	–	l
15	H ₂ O	single	water vapor	•	•	–	–	–	s
16	OH	single	hydroxyl radical	•	–	–	•	–	s
17	HO ₂	single	hydroperoxy radical	•	–	–	•	–	s
18	H ₂ O ₂	single	hydrogen peroxide	•	•	–	•	•	s
19	CH ₄	single	methane	•	•	•	–	–	l
20	CH ₃ O ₂	single	methyl peroxy radical	•	–	–	–	–	s
21	CH ₃ O	single	methoxy radical	•	–	–	–	–	s
22	CH ₃ OOH	single	methyl hydroperoxide	•	•	–	•	•	s
23	CH ₃ OH	single	methanol	•	•	•	•	•	s
24	CH ₂ O	single	formaldehyde	•	•	–	•	•	s
25	CO	single	carbon oxide	•	•	•	•	–	l
26	CO2FF	single	carbon dioxide ^k (inert tracer)	–	•	•	–	–	l
27	C ₂ H ₆	single	ethane	•	•	•	•	–	s
28	C ₂ H ₅ O ₂	single	ethyl peroxy radical	•	–	–	–	–	s
29	C ₂ H ₅ OH	single	ethanol and higher alcohols	•	•	–	•	•	s

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Table 1. Continued.

#	tracer	family ^a	description	ch. ^b	a/c ^c	em. ^d	d.d. ^e	w.s. ^f	τ ^g
30	C ₂ H ₅ OOH	single	ethyl hydroperoxide	•	•	–	•	•	S
31	CH ₃ CHO	single	acetaldehyde	•	•	–	•	•	S
32	CH ₃ CO ₃	single	peroxy acetyl radical	•	–	–	–	–	S
33	CH ₃ COOH	single	acetic acid and higher organic acids	•	•	–	•	•	S
34	CH ₃ C(O)OOH	single	paracetic acid	•	•	–	•	•	S
35	C ₃ H ₈	single	propane	•	•	•	•	–	S
36	C ₃ H ₇ O ₂	single	propyl peroxy radical	•	–	–	–	–	S
37	C ₃ H ₇ OOH	single	propyl hydroperoxide	•	•	–	•	•	S
38	CH ₃ COCH ₃	single	acetone	•	•	•	•	•	S
39	PROPAO ₂	single	CH ₃ COCH ₂ O ₂	•	–	–	–	–	S
40	PROPAOOH	single	CH ₃ COCH ₂ OOH	•	•	–	•	•	S
41	CH ₃ COCHO	single	methylglyoxal	•	•	–	•	•	S
42	ALKAN	single	C ₄ - and higher alkanes	•	•	•	–	–	S
43	ALKANO ₂	single	C ₄ - and higher alkanyl peroxy radicals	•	•	–	–	–	S
44	ALKANOOH	single	C ₄ - and higher alkanyl hydroperoxides	•	•	–	•	•	S
45	MEK	single	methyl ethyl ketone (CH ₃ CH ₂ COCH ₃)	•	•	•	•	•	S
46	MEKO ₂	single	peroxy radical from MEK	•	–	–	–	–	S
47	MEKOOH	single	hydroperoxide from MEK	•	•	–	•	•	S
48	C ₂ H ₄	single	ethene	•	•	•	•	–	S
49	C ₃ H ₆	single	propene	•	•	•	•	–	S
50	PROPEO ₂	single	hydroxy propyl peroxy radical	•	–	–	–	–	S
51	PROPEOOH	single	hydroxy propyl hydroperoxide	•	•	–	•	•	S
52	C ₂ H ₂	single	ethine	•	•	•	•	–	S
53	ALKEN	single	C ₄ - and higher alkenes	•	•	•	–	–	S
54	ALKENO ₂	single	C ₄ - and higher alkenyl peroxy radicals	•	–	–	–	–	S
55	ALKENOOH	single	C ₄ - and higher alkenyl hydroperoxides	•	•	–	•	•	S
56	AROM	single	aromatic compounds	•	•	•	–	–	S
57	AROMO ₂	single	peroxy radicals from AROM	•	–	–	–	–	S
58	AROMOOH	single	hydroperoxides from AROM	•	•	–	•	•	S

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Table 1. Continued.

#	tracer	family ^a	description	ch. ^b	a/c ^c	em. ^d	d.d. ^e	w.s. ^f	τ ^g
59	ISOP	single	isoprene	•	•	•	•	–	s
60	ISOPO2	single	peroxy radical from isoprene	•	–	–	–	–	s
61	ISOPNO3	single	nitrate radical adduction to isoprene	•	–	–	–	–	s
62	MVK	single	methyl vinyl ketone (CH ₂ CHCOCH ₃)	•	•	•	•	•	s
63	MACR	single	methacrolein (CH ₂ CCH ₃ CHO)	•	•	–	•	–	s
64	MACRO2	single	peroxy radical from methacrolein	•	–	–	–	–	s
65	MACROOH	single	hydroperoxide from methacrolein	•	•	–	•	•	s
66	MACO3	single	carboxy radical from methacrolein	•	–	–	–	–	s
67	APIN	single	α -pinene and other terpenes	•	•	•	•	–	s
68	APINO2	single	peroxy radicals from APIN	•	–	–	–	–	s
69	APINO3	single	ozone adduction to terpenes	•	–	–	–	–	s
70	PCHO	single	aldehydes from APIN	•	•	–	•	•	s
71	PCO3	single	carboxy radicals from APIN	•	–	–	–	–	s
72	ONITU	single	unreactive organic nitrates	•	•	–	•	•	s
73	ONITUO2	single	peroxy radicals from ONITU + OH	•	–	–	–	–	s
74	ONITR	single	reactive organic nitrates	•	•	–	•	•	s
75	PAN	single	peroxy acetyl nitrate	•	•	–	•	•	s
76	MPAN	single	PAN analog from MACR	•	•	–	•	–	s
77	APINPAN	single	PAN analog from APIN	•	•	–	•	–	s
78	PCO3PAN	single	PAN analog from PCHO	•	•	–	•	–	s
79	XO2	single	counter for organic peroxy radicals	•	–	–	–	–	s
80	XOOH	single	counter for organic hydroperoxides	•	•	–	•	•	s
81	MCF	single	methyl chloroform (CH ₃ CCl ₃)	•	•	•	–	–	l
82	²²² Rn	single	radon	•	•	•	–	–	l
83	²¹⁰ Pb	single	lead	•	•	–	•	•	l

^aused only in transport ^eSpecies is subject to dry deposition ⁱmarker for stratospheric O₃ intrusion

^bSpecies is subject to chemistry ^fSpecies is subject to wet scavenging ^jNO_x is emitted as NO

^cSpecies is subject to advection/ convection ^gs: short-lived, l: long-lived ^kFossil fuel combustion

^dSpecies has a surface source ^hnot subject to O₃ photochemistry

Table 2. Photolysis reactions in LMDz-INCA.

#	reaction	refs.
j1	$O_2 + hv \rightarrow 2 * O(^3P)$	1
j2	$O_3 + hv \rightarrow O(^1D) + O_2$	1
j3	$O_3 + hv \rightarrow O(^3P) + O_2$	1
j4	$H_2 + hv \rightarrow OH + H$	1
j5	$HO_2 + hv \rightarrow 2 * OH$	1
j6	$NO + hv \rightarrow O(^3P) + N$	2
j7	$NO_2 + hv \rightarrow NO + O(^3P)$	1
j8	$NO_3 + hv \rightarrow NO + O_2$	1
j9	$NO_3 + hv \rightarrow NO_2 + O(^3P)$	1
j10	$N_2O + hv \rightarrow O(^1D) + N_2$	1
j11	$N_2O_5 + hv \rightarrow NO_2 + NO_3$	1
j12	$HNO_2 + hv \rightarrow NO + OH$	1
j13	$HNO_3 + hv \rightarrow NO_2 + OH$	1
j14	$HNO_4 + hv \rightarrow NO_2 + HO_2$	1
j15	$CH_3OOH + hv \rightarrow CH_2O + OH + H$	1
j16	$CH_2O + hv \rightarrow CO + 2 * H$	1
j17	$CH_2O + hv \rightarrow CO + H_2$	1
j18	$CH_3CHO + hv \rightarrow CH_3O_2 + CO + HO_2$	1
j19	$MACR + hv \rightarrow 0.67 * HO_2 + 0.33 * MCO_3 +$ $0.67 * CH_2O + 0.67 * CO +$ $0.67 * CH_3CO_3 + 0.33 * OH$	1
j20	$PCHO + hv \rightarrow HO_2 + CO + XO_2$	3
j21	$CH_3COCH_3 + hv \rightarrow CH_3CO_3 + CH_3O_2$	1
j22	$CH_3COCHO + hv \rightarrow CH_3CO_3 + CO + HO_2$	1
j23	$MEK + hv \rightarrow CH_3CO_3 + C_2H_5O_2$	4
j24	$MVK + hv \rightarrow 0.3 * CH_3CO_3 + 0.7 * C_3H_6 +$ $0.7 * CO + 0.3 * CH_3O_2$	1

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Table 2. Continued.

#	reaction	refs.
j25	ONITU + $h\nu$ → NO ₂ + 0.512 * MEK + 0.33 * CH ₃ COCH ₃ + 0.346 * C ₂ H ₅ O ₂ + 0.653 * HO ₂	8
j26	ONITR + $h\nu$ → HO ₂ + CO + NO ₂ + CH ₂ O	3
j27	PAN + $h\nu$ → CH ₃ CO ₃ + NO ₂	1
j28	MPAN + $h\nu$ → MCO ₂ + NO ₂	5
j29	APINPAN + $h\nu$ → APINO ₃ + NO ₂	5
j30	PCO ₃ PAN + $h\nu$ → PCO ₃ + NO ₂	5
j31	CH ₃ C(O)OOH + $h\nu$ → CH ₃ O ₂ + OH + CO ₂	7
j32	C ₂ H ₅ OOH + $h\nu$ → CH ₃ CHO + HO ₂ + OH	6
j33	C ₃ H ₇ OOH + $h\nu$ → 0.218 * CH ₃ CHO + OH + HO ₂ + 0.782 * CH ₃ COCH ₃	6
j34	PROP ₃ OOH + $h\nu$ → CH ₂ O + CH ₃ CO ₃ + OH	6
j35	PROPEOOH + $h\nu$ → CH ₃ CHO + CH ₂ O + HO ₂ + OH	6
j36	ALKANOOH + $h\nu$ → 0.562 * HO ₂ + 0.334 * XO ₂ + 0.358 * CH ₃ COCH ₂ + OH + 0.514 * MEK + 0.006 * CH ₂ O + 0.454 * CH ₃ CHO + 0.1 * CH ₃ O ₂	6
j37	ALKENOOH + $h\nu$ → OH + HO ₂ + 0.165 * CH ₂ O + 0.68 * CH ₃ CHO + 0.155 * CH ₃ COCH ₃	6
j38	AROMOOH + $h\nu$ → OH + 0.423 * CH ₃ COCH ₃ + 1.658 * HO ₂ + 0.658 * CO + 0.658 * CH ₃ CO ₃	6
j39	MACROOH + $h\nu$ → OH + 0.84 * CO + HO ₂ + 0.16 * CH ₂ O + CH ₃ COCHO	6
j40	MACROOH + $h\nu$ → 1.16 + HO ₂ + CO + 0.84 * OH + 0.84 * CH ₃ COCHO + 0.16 * PROP ₃ OOH	3

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Table 2. Continued.

#	reaction	refs.
j41	MEKOOH + $h\nu$ → OH + 0.93 * CH ₃ CHO + 0.6 * CH ₃ CO ₃ + 0.07 * CH ₂ O + 0.2 * MEK	6
j42	XOOH + $h\nu$ → OH	6
j43	MCF + $h\nu$ → (no products considered)	1

references:
 1, Madronich and Flocke (1998); 2, Brasseur and Solomon (1987)
 3, $J = J(\text{CH}_3\text{CHO})$; 4, $J = 1.7 \times J(\text{CH}_3\text{COCH}_3)$
 5, $J = J(\text{PAN})$; 6, $J = J(\text{CH}_3\text{OOH})$; 7, $J = 0.28 \times J(\text{H}_2\text{O}_2)$
 8, $J(\text{ONITU}) = J(\text{CH}_3\text{ONO}_2)$

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Table 3. Heterogenous reactions included in LMDz-INCA.

#	Reaction	Reaction probability
h1	$\text{N}_2\text{O}_5 \rightarrow 2 \text{HNO}_3$	$\gamma_{200 \text{ K}} = 0.185,$ $\gamma_{300 \text{ K}} = 0.03$
h2	$\text{NO}_3 \rightarrow \text{HNO}_3$	$\gamma = 0.1$
h3	$\text{NO}_2 \rightarrow 0.5 \text{HNO}_3 + 0.5 \text{HNO}_2$	$\gamma = 0.1$
h4	$\text{HO}_2 \rightarrow 0.5 \text{H}_2\text{O}_2 + 0.5 \text{O}_2$	$\gamma = 1.0 \times 10^{-3}$

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Table 4. Thermochemical reactions in LMDz-INCA.

#	reaction	rate coefficient	refs.
k1	$O(^3P)+O_2+M \rightarrow O_3+M$	$6.0 \times 10^{-34} (300/T)^{2.3}$	2
k2	$O(^3P)+O(^3P)+M \rightarrow O_2+M$	$4.23 \times 10^{-28} T^{-2.0}$	14, 15
k3	$O(^3P)+O_3 \rightarrow 2 O_2$	$8.0 \times 10^{-12} \exp(-2060/T)$	1
k4	$O(^1D)+O_3 \rightarrow 2 O_2$	1.2×10^{-10}	2
k5	$O(^1D)+N_2 \rightarrow O(^3P)+N_2$	$1.8 \times 10^{-11} \exp(107/T)$	1
k6	$O(^1D)+O_2 \rightarrow O(^3P)+O_2$	$3.2 \times 10^{-11} \exp(67/T)$	1
k7	$O(^1D)+H_2O \rightarrow 2 OH$	2.2×10^{-10}	2
k8	$O(^1D)+H_2 \rightarrow OH+H$	1.1×10^{-10}	1, 2
k9	$O(^1D)+N_2O \rightarrow 2 NO$	6.7×10^{-11}	2
k10	$O(^1D)+N_2O \rightarrow N_2+O_2$	4.4×10^{-11}	1
k11	$O_2+N \rightarrow NO+O(^3P)$	$1.5 \times 10^{-11} \exp(-3600/T)$	2
k12	$O_2+H+M \rightarrow HO_2+M$	$k_0 = 5.69 \times 10^{-32} (T/298)^{-1.6}$ $k_\infty = 7.5 \times 10^{-11}$ $F_c = 0.6$	2
k13	$O_3+H \rightarrow OH+O_2$	$1.0 \times 10^{-10} \exp(-367/T)$	5
k14	$OH+O(^3P) \rightarrow H+O_2$	$2.2 \times 10^{-11} \exp(120/T)$	2
k15	$OH+O_3 \rightarrow HO_2+O_2$	$1.9 \times 10^{-12} \exp(-1000/T)$	1
k16	$OH+H_2 \rightarrow H_2O+H$	$5.5 \times 10^{-12} \exp(-2000/T)$	2
k17	$OH+OH \rightarrow H_2O+O(^3P)$	$4.2 \times 10^{-12} \exp(-240/T)$	2
k18	$OH+HO_2 \rightarrow H_2O+O_2$	$4.8 \times 10^{-11} \exp(250/T)$	1, 2
k19	$OH+H_2O_2 \rightarrow H_2O+HO_2$	$2.9 \times 10^{-12} \exp(-160/T)$	1, 2
k20	$OH+HNO_2 \rightarrow H_2O+NO_2$	$1.8 \times 10^{-11} \exp(-390/T)$	2
k21	$OH+HNO_3 \rightarrow H_2O+NO_3$	$k_1 = 7.2 \times 10^{-15} \exp(785/T)$ $k_2 = 4.1 \times 10^{-16} \exp(1440/T)$ $k_3 = 1.9 \times 10^{-33} \exp(725/T) [M]$	13
			$k = k_1 + \frac{k_3}{1 + \frac{k_3}{k_2}}$
k22	$OH+HNO_4 \rightarrow H_2O+NO_2+O_2$	$1.3 \times 10^{-12} \exp(380/T)$	2
k23	$HO_2+O(^3P) \rightarrow OH+O_2$	$3.0 \times 10^{-11} \exp(200/T)$	2
k24	$HO_2+O_3 \rightarrow OH+2 O_2$	$1.1 \times 10^{-14} \exp(-500/T)$	2
k25	$HO_2+H \rightarrow 2 OH$	7.2×10^{-11}	1
k26	$HO_2+H \rightarrow H_2+O_2$	5.6×10^{-12}	1
k27	$HO_2+H \rightarrow H_2O+O(^3P)$	2.4×10^{-12}	1

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k28	HO ₂ +HO ₂ → H ₂ O ₂ +O ₂	$k_0 = 2.3 \times 10^{-13} (600/T)$ $k_\infty = 1.7 \times 10^{-33} [M] \exp(1000/T)$ $F_c = 1.0 + 1.4 \times 10^{-21} [H_2O] \exp(2200/T)$ $k = (k_0 + k_\infty) F_c$	14,15
k29	NO+O ₃ → NO ₂ +O ₂	$1.8 \times 10^{-12} \exp(-1370/T)$	1
k30	NO+OH+M → HNO ₂ +M	$k_0 = 7.01 \times 10^{-31} (T/298)^{-2.6}$ $k_\infty = 3.6 \times 10^{-11} (T/298)^{-0.1}$ $F_c = 0.6$	2
k31	NO+HO ₂ → NO ₂ +OH	$3.5 \times 10^{-12} \exp(250/T)$	2
k32	NO+N → N ₂ +O(³ P)	$2.1 \times 10^{-11} \exp(100/T)$	2
k33	NO ₂ +O(³ P) → NO+O ₂	$6.5 \times 10^{-12} \exp(120/T)$	1,2
k34	NO ₂ +O(³ P)+M → NO ₃ +M	$k_0 = 9.0 \times 10^{-32} (T/298)^{-2.0}$ $k_\infty = 2.2 \times 10^{-11}$ $F_c = 0.6$	2
k35	NO ₂ +O ₃ → NO ₃ +O ₂	$1.2 \times 10^{-13} \exp(-2450/T)$	2
k36	NO ₂ +H → OH+NO	$4.0 \times 10^{-10} \exp(-340/T)$	2
k37	NO ₂ +OH+M → HNO ₃ +M	$k_0 = 2.6 \times 10^{-30} (T/298)^{-2.9}$ $k_\infty = 7.5 \times 10^{-11} (T/298)^{-0.6}$ $F_c = 0.6$	1
k38	NO ₂ +HO ₂ +M → HNO ₄ +M	$k_0 = 1.8 \times 10^{-31} (T/298)^{-3.2}$ $k_\infty = 4.7 \times 10^{-12} (T/298)^{-1.4}$ $F_c = 0.6$	13
k39	HNO ₄ +M → NO ₂ +HO ₂ +M	$k_0 = 1.8 \times 10^{-31} (T/298)^{-3.2}$ $k_\infty = 4.7 \times 10^{-12} (T/298)^{-1.4}$ $k_{inv} = 2.1 \times 10^{-27} \exp(10900/T)$ $F_c = 0.6$	13
k40	NO ₂ +NO ₃ +M → N ₂ O ₅ +M	$k_0 = 2.0 \times 10^{-30} (T/298)^{-4.4}$ $k_\infty = 1.4 \times 10^{-12} (T/298)^{-0.7}$ $F_c = 0.6$	14,15
k41	N ₂ O ₅ +M → NO ₂ +NO ₃ +M	$k_0 = 2.0 \times 10^{-30} (T/298)^{-4.4}$ $k_\infty = 1.4 \times 10^{-12} (T/298)^{-0.7}$ $k_{inv} = 3.0 \times 10^{-27} \exp(10991/T)$ $F_c = 0.6$	14,15
k42	NO ₃ +NO → 2NO ₂	$1.8 \times 10^{-11} \exp(110/T)$	1

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k43	$\text{NO}_3 + \text{HO}_2 \rightarrow 0.4\text{HNO}_3 + 0.6\text{OH} + 0.6\text{NO}_2$	$2.3 \times 10^{-12} \exp(170/T)$	6
k44	$\text{CH}_3\text{CCl}_3 + \text{OH} \rightarrow \text{H}_2\text{O}$	$1.8 \times 10^{-12} \exp(-1550/T)$	2
k45	$\text{CH}_4 + \text{OH} \rightarrow \text{CH}_3\text{O}_2 + \text{H}_2\text{O}$	$2.45 \times 10^{-12} \exp(-1775/T)$	2
k46	$\text{CH}_4 + \text{O}(^1\text{D}) \rightarrow \text{CH}_3\text{O}_2 + \text{OH}$	2.25×10^{-10}	7
k47	$\text{CH}_4 + \text{O}(^1\text{D}) \rightarrow \text{CH}_2\text{O} + \text{H}_2$	1.65×10^{-11}	1,2,8
k48	$\text{CH}_4 + \text{O}(^1\text{D}) \rightarrow \text{CH}_3\text{OH}$	4.98×10^{-11}	9
k49	$\text{CH}_3\text{O}_2 + \text{NO} \rightarrow \text{CH}_3\text{O} + \text{NO}_2$	$3.0 \times 10^{-12} \exp(280/T)$	2
k50	$\text{CH}_3\text{O}_2 + \text{NO}_3 \rightarrow \text{CH}_3\text{O} + \text{NO}_2 + \text{O}_2$	3.1×10^{-12}	10
k51	$\text{CH}_3\text{O}_2 + \text{HO}_2 \rightarrow \text{CH}_3\text{OOH} + \text{O}_2$	$3.8 \times 10^{-13} \exp(800/T)$	1,2
k52	$\text{CH}_3\text{O}_2 + \text{CH}_3\text{O}_2 \rightarrow \text{CH}_3\text{OH} + \text{CH}_2\text{O} + \text{O}_2$	$1.5 \times 10^{-13} \exp(190/T)$	2,11
k53	$\text{CH}_3\text{O}_2 + \text{CH}_3\text{O}_2 \rightarrow 2\text{CH}_3\text{O} + \text{O}_2$	$1.0 \times 10^{-13} \exp(190/T)$	2,11
k54	$\text{CH}_3\text{O} + \text{O}_2 \rightarrow \text{CH}_2\text{O} + \text{HO}_2$	$3.9 \times 10^{-14} \exp(-900/T)$	2
k55	$\text{CH}_3\text{O} + \text{NO}_2 \rightarrow \text{CH}_2\text{O} + \text{HNO}_2$	$1.1 \times 10^{-11} \exp(-1200/T)$	2
k56	$\text{CH}_3\text{OH} + \text{OH} \rightarrow \text{CH}_2\text{O} + \text{HO}_2 + \text{H}_2\text{O}$	$3.1 \times 10^{-12} \exp(-360/T)$	1
k57	$\text{CH}_3\text{OOH} + \text{OH} \rightarrow \text{CH}_2\text{O} + \text{OH} + \text{H}_2\text{O}$	$1.0 \times 10^{-12} \exp(190/T)$	1
k58	$\text{CH}_3\text{OOH} + \text{OH} \rightarrow \text{CH}_3\text{O}_2 + \text{H}_2\text{O}$	$1.9 \times 10^{-12} \exp(190/T)$	1
k59	$\text{CH}_2\text{O} + \text{OH} \rightarrow \text{CO} + \text{HO}_2 + \text{H}_2\text{O}$	$8.59 \times 10^{-12} \exp(20/T)$	1
k60	$\text{CH}_2\text{O} + \text{NO}_3 \rightarrow \text{CO} + \text{HO}_2 + \text{HNO}_3$	5.8×10^{-16}	1
k61	$\text{CH}_2\text{O} + \text{O}(^3\text{P}) \rightarrow \text{CO} + \text{HO}_2 + \text{OH}$	$3.4 \times 10^{-11} \exp(-1600/T)$	2,12
k62	$\text{CO} + \text{OH} \rightarrow \text{CO}_2 + \text{H}$	$1.57 \times 10^{-13} + 3.54 \times 10^{-33} [\text{M}]$	16
k63	$\text{C}_2\text{H}_6 + \text{OH} \rightarrow \text{C}_2\text{H}_5\text{O}_2 + \text{H}_2\text{O}$	$1.52 \times 10^{-17} \exp(-498/T) \text{ T}^2$	13
k64	$\text{C}_2\text{H}_5\text{O}_2 + \text{NO} \rightarrow \text{CH}_3\text{CHO} + \text{HO}_2 + \text{NO}_2$	$2.7 \times 10^{-12} \exp(360/T)$	13
k65	$\text{C}_2\text{H}_5\text{O}_2 + \text{NO}_3 \rightarrow \text{CH}_3\text{CHO} + \text{HO}_2 + \text{NO}_2 + \text{O}_2$	2.4×10^{-12}	13
k66	$\text{C}_2\text{H}_5\text{O}_2 + \text{HO}_2 \rightarrow \text{C}_2\text{H}_5\text{OOH} + \text{O}_2$	$4.4 \times 10^{-13} \exp(900/T)$	13
k67	$\text{C}_2\text{H}_5\text{O}_2 + \text{CH}_3\text{O}_2 \rightarrow 0.74\text{CH}_2\text{O} + 0.74\text{CH}_3\text{CHO} + 0.96\text{HO}_2 + 0.26\text{CH}_3\text{OH} + 0.26\text{C}_2\text{H}_5\text{OH}$	2.0×10^{-13}	13
k68	$\text{C}_2\text{H}_5\text{O}_2 + \text{C}_2\text{H}_5\text{O}_2 \rightarrow 1.63\text{CH}_3\text{CHO} + 1.26\text{HO}_2 + 0.37\text{C}_2\text{H}_5\text{OH}$	$9.8 \times 10^{-14} \exp(100/T)$	13
k69	$\text{C}_2\text{H}_5\text{OOH} + \text{OH} \rightarrow \text{C}_2\text{H}_5\text{O}_2 + \text{HO}_2$	$1.9 \times 10^{-12} \exp(190/T)$	13
k70	$\text{C}_2\text{H}_5\text{OOH} + \text{OH} \rightarrow \text{CH}_3\text{CHO} + \text{OH} + \text{H}_2\text{O}$	$7.69 \times 10^{-17} \exp(253/T) \text{ T}^2$	13
k71	$\text{C}_2\text{H}_5\text{OH} + \text{OH} \rightarrow \text{CH}_3\text{CHO} + \text{HO}_2 + \text{H}_2\text{O}$	$6.18 \times 10^{-18} \exp(532/T) \text{ T}^2$	13
k72	$\text{C}_3\text{H}_8 + \text{OH} \rightarrow \text{C}_3\text{H}_7\text{O}_2 + \text{H}_2\text{O}$	$1.55 \times 10^{-17} \exp(-61/T) \text{ T}^2$	13
k73	$\text{C}_3\text{H}_7\text{O}_2 + \text{NO} \rightarrow 0.72\text{CH}_3\text{COCH}_3 + 0.94\text{NO}_2 + 0.22\text{CH}_3\text{CHO} + 0.94\text{HO}_2 + 0.06\text{ONITU}$	$2.7 \times 10^{-12} \exp(360/T)$	13
k74	$\text{C}_3\text{H}_7\text{O}_2 + \text{NO}_3 \rightarrow 0.234\text{CH}_3\text{CHO} + \text{NO}_2 + \text{HO}_2 + 0.766\text{CH}_3\text{COCH}_3$	2.4×10^{-12}	13
k75	$\text{C}_3\text{H}_7\text{O}_2 + \text{HO}_2 \rightarrow \text{C}_3\text{H}_7\text{OOH} + \text{O}_2$	$1.9 \times 10^{-13} \exp(1300/T)$	13

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Table 4. Continued.

#	reaction	rate coefficient	refs.	
k76	$C_3H_7O_2 + CH_3O_2 \rightarrow$	$0.128CH_3COCH_3 + 0.78HO_2 +$ $0.695CH_2O + 0.305CH_3OH +$ $0.567CH_3CHO$	5.18×10^{-12}	13
k77	$C_3H_7OOH + OH \rightarrow$	$C_3H_7O_2 + H_2O$	$1.9 \times 10^{-12} \exp(190/T)$	13
k78	$C_3H_7OOH + OH \rightarrow$	$CH_3CHO + OH + H_2O$	$1.67 \times 10^{-17} \exp(253/T)T^2$	13
k79	$CH_3COCH_3 + OH \rightarrow$	$PROPAO_2 + H_2O$	$2.81 \times 10^{-12} \exp(-760/T)$	1
k80	$PROPAO_2 + NO \rightarrow$	$CH_3CO_3 + CH_2O + NO_2$	$2.7 \times 10^{-12} \exp(360/T)$	13
k81	$PROPAO_2 + NO_3 \rightarrow$	$CH_3CO_3 + CH_2O + NO_2 + O_2$	2.4×10^{-12}	13
k82	$PROPAO_2 + HO_2 \rightarrow$	$PROPAOOH + O_2$	$1.9 \times 10^{-13} \exp(1300/T)$	13
k83	$PROPAO_2 + CH_3O_2 \rightarrow$	$1.31CH_2O + 0.23CH_3OH +$ $0.23CH_3COCHO + 0.54HO_2 +$ $0.54CH_3CO_3$	3.8×10^{-12}	13
k84	$PROPAOOH + OH \rightarrow$	$PROPAO_2 + H_2O$	$1.9 \times 10^{-12} \exp(190/T)$	13
k85	$PROPAOOH + OH \rightarrow$	$CH_3COCHO + OH + H_2O$	$4.69 \times 10^{-17} \exp(253/T)T^2$	13
k86	$C_2H_4 + OH + M \rightarrow$	$0.667PROPEO_2 + M$	$k_0 = 1.0 \times 10^{-28} (T/298)^{-0.8}$ $k_{\infty} = 8.79 \times 10^{-12}$ $F_c = 0.7$	2
k87	$C_2H_4 + O_3 \rightarrow$	$CH_2O + 0.46CO + 0.16HO_2 +$ $0.08OH + 0.17CO_2$	$1.2 \times 10^{-14} \exp(-2630/T)$	2
k88	$C_3H_6 + OH + M \rightarrow$	$PROPEO_2 + M$	$k_0 = 2.94 \times 10^{-27} (T/298)^{-3.0}$ $k_{\infty} = 2.775 \times 10^{-11} (T/298)^{-1.3}$ $F_c = 0.5$	13
k89	$C_3H_6 + O_3 \rightarrow$	$0.63248CH_2O + 0.341CH_3O_2 +$ $0.0868CH_4 + 0.4166CO +$ $0.0124CH_3OH + 0.2096HO_2 +$ $0.2474OH + 0.38CH_3CHO + 0.2754CO_2$	$6.51 \times 10^{-15} \exp(-1900/T)$	2
k90	$PROPEO_2 + NO \rightarrow$	$CH_3CHO + CH_2O + HO_2 + NO_2$	$2.7 \times 10^{-12} \exp(360/T)$	13
k91	$PROPEO_2 + NO_3 \rightarrow$	$CH_3CHO + CH_2O + HO_2 + NO_2 + O_2$	2.4×10^{-12}	13
k92	$PROPEO_2 + HO_2 \rightarrow$	$PROPEOOH + O_2$	$1.9 \times 10^{-13} \exp(1300/T)$	13
k93	$PROPEO_2 + CH_3O_2 \rightarrow$	$0.305CH_3OH + 0.78HO_2 +$ $1.085CH_2O + 0.39CH_3CHO +$ $0.305CH_3COCHO$	7.583×10^{-13}	13
k94	$PROPEOOH + OH \rightarrow$	$PROPEO_2 + H_2O$	$1.9 \times 10^{-12} \exp(190/T)$	13
k95	$PROPEOOH + OH \rightarrow$	$CH_3COCHO + OH + H_2O$	$2.35 \times 10^{-17} \exp(696/T)T^2$	13
k96	$PROPEOOH + OH \rightarrow$	$CH_3COCHO + OH + H_2O$	$2.69 \times 10^{-17} \exp(253/T)T^2$	13
k97	$PROPEOOH + OH \rightarrow$	$CH_3CHO + HO_2 + H_2O$	$1.26 \times 10^{-17} \exp(253/T)T^2$	13
k98	$PROPEOOH + OH \rightarrow$	$PROPAOOH + HO_2 + H_2O$	$3.19 \times 10^{-18} \exp(696/T)T^2$	13

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#	reaction	rate coefficient	refs.
k99	CH ₃ CHO+OH → CH ₃ CO ₃ +H ₂ O	5.6×10 ⁻¹² exp(270/T)	2
k100	CH ₃ CHO+NO ₃ → CH ₃ CO ₃ +HNO ₃	1.4×10 ⁻¹² exp(-1860/T)	13
k101	CH ₃ CO ₃ +NO → CH ₃ O ₂ +NO ₂ +CO ₂	5.3×10 ⁻¹² exp(360/T)	13
k102	CH ₃ CO ₃ +NO ₂ +M → PAN+M	k ₀ = 2.7×10 ⁻²⁸ (T/298) ^{-7.1} k _∞ = 1.2×10 ⁻¹¹ (T/298) ^{-0.9} F _c = 0.3	13
k103	PAN+M → CH ₃ CO ₃ +NO ₂ +M	k ₀ = 5.0×10 ⁻² exp(-12875/T) k _∞ = 2.2×10 ⁺¹⁶ exp(-13435/T) F _c = 0.27	13
k104	CH ₃ CO ₃ +NO ₃ → CH ₃ O ₂ +NO ₂ +CO ₂ +O ₂	5.0×10 ⁻¹²	13
k105	CH ₃ CO ₃ +HO ₂ → 0.3O ₃ +0.3CH ₃ OOH+0.7O ₂ +0.7CH ₃ C(O)OOH	4.3×10 ⁻¹³ exp(1040/T)	13
k106	CH ₃ CO ₃ +CH ₃ O ₂ → CH ₂ O+0.86CH ₃ O ₂ +0.86HO ₂ +0.86CO ₂ +O ₂ +0.14CH ₃ COOH	1.3×10 ⁻¹² exp(640/T)	13
k107	CH ₃ CO ₃ +CH ₃ CO ₃ → 2CH ₃ O ₂ +2CO ₂	2.3×10 ⁻¹² exp(530/T)	13
k108	CH ₃ C(O)OOH+OH → CH ₃ CO ₃ +H ₂ O	1.9×10 ⁻¹² exp(190/T)	13
k109	C ₂ H ₂ +OH+M → 0.36CO+0.64CH ₃ COCHO+0.36HO ₂ +0.65OH+M	k ₀ = 5.01×10 ⁻³⁰ (T/298) ^{-1.5} k _∞ = 9.0×10 ⁻¹³ (T/298) ^{2.0} F _c = 0.62	1
k110	ISOP+OH → ISOPO ₂	2.89×10 ⁻¹¹ exp(335/T)	1,2
k111	ISOP+O ₃ → 0.42MACR+0.16MVK+0.05C ₃ H ₆ +0.18OH+0.09HO ₂ +0.42CH ₂ O+0.27CO+0.07H ₂ +0.15CO ₂	9.36×10 ⁻¹⁵ exp(-1913/T)	13
k112	ISOP+NO ₃ → ISOPNO ₃	3.03×10 ⁻¹² exp(-446/T)	13
k113	ISOPO ₂ +NO → 0.12ONITR+0.88NO ₂ +0.76HO ₂ +0.608CH ₂ O+0.404MACR+0.354MVK+0.12XO ₂	2.7×10 ⁻¹² exp(360/T)	13
k114	ISOPO ₂ +NO ₃ → 0.864HO ₂ +NO ₂ +0.69CH ₂ O+0.46MACR+0.403MVK+0.136XO ₂	2.4×10 ⁻¹²	13
k115	ISOPO ₂ +HO ₂ → 0.867HO ₂ +0.739CH ₂ O+0.506MACR+0.429MVK+0.133XO ₂ +XOOH	1.9×10 ⁻¹³ exp(1300/T)	13
k116	ISOPO ₂ +CH ₃ O ₂ → 0.305CH ₃ OH+0.703HO ₂ +0.91CH ₂ O+0.137XO ₂ +0.351MACR+0.205MVK	1.33×10 ⁻¹²	13

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k117	ISOPO ₂ +CH ₃ CO ₃ →	0.275CH ₃ COOH+0.58HO ₂ + 0.725CO+0.725CH ₃ O ₂ + 0.198XO+20.397CH ₂ O+ 0.504MACR+0.296MVK	7.96×10 ⁻¹² 13
k118	ISOPNO ₃ +NO →	1.206NO ₂ +0.794HO ₂ + 0.072CH ₂ O+0.167MACR+ 0.039MVK+0.794ONITR	2.7×10 ⁻¹² exp(360/T) 13
k119	ISOPNO ₃ +NO ₃ →	1.206NO ₂ +0.794HO ₂ + 0.072CH ₂ O+0.167MACR+ 0.039MVK+0.794ONITR	2.4×10 ⁻¹² 13
k120	ISOPNO ₃ +HO ₂ →	0.206NO ₂ +0.794HO ₂ + 0.008CH ₂ O+0.167MACR+ 0.039MVK+0.794ONITR+XOOH	1.9×10 ⁻¹³ exp(1300/T) 13
k121	ISOPNO ₃ +CH ₃ O ₂ →	0.305CH ₃ OH+0.711HO ₂ + 0.697CH ₂ O+0.06NO ₂ + 0.059MACR+0.001MVK+ 0.635ONITR	1.749×10 ⁻¹² 13
k122	APIN+OH →	APINO ₂	1.08×10 ⁻¹¹ exp(444/T) 13
k123	APIN+O ₃ →	0.56OH+0.56APINO ₃	1.1615×10 ⁻¹⁵ exp(-732/T) 13
k124	APIN+NO ₃ →	APINO ₂ +NO ₂	1.19×10 ⁻¹² exp(490/T) 13
k125	APINO ₂ +NO →	PCHO+NO ₂ +HO ₂	2.7×10 ⁻¹² exp(360/T) 13
k126	APINO ₂ +NO ₃ →	PCHO+NO ₂ +HO ₂ +O ₂	2.4×10 ⁻¹² 13
k127	APINO ₂ +HO ₂ →	PCHO+HO ₂ +XOOH	1.9×10 ⁻¹³ exp(1300/T) 13
k128	APINO ₂ +CH ₃ O ₂ →	0.305CH ₃ OH+0.695APINO ₃ +HO ₂	1.22×10 ⁻¹³ 13
k129	APINO ₂ +CH ₃ CO ₃ →	0.725CO ₂ +0.725CH ₃ O ₂ + 0.725PCHO+0.725HO ₂ + 0.275CH ₃ COOH	7.37×10 ⁻¹³ 13
k130	APINO ₃ +NO →	NO ₂ +CO ₂ +CH ₃ COCH ₃ + 2PROPEO ₂	5.3×10 ⁻¹² exp(360/T) 13
k131	APINO ₃ +NO ₂ +M →	APINPAN+M	k ₀ = 2.7×10 ⁻²⁸ (T/298) ^{-7.1} k _∞ = 1.2×10 ⁻¹¹ (T/298) ^{-0.9} F _c = 0.3 13
k132	APINPAN+M →	APINO ₃ +NO ₂ +M	k ₀ = 4.0×10 ⁻³ exp(-12100/T) k _∞ = 5.4×10 ⁺¹⁶ exp(-13830/T) F _c = 0.3 14,15
k133	APINO ₃ +NO ₃ →	NO ₂ +CO ₂ +CH ₃ COCH ₃ + 2PROPEO ₂ +O ₂	5.0×10 ⁻¹² 13
k134	APINO ₃ +HO ₂ →	0.3O ₃ +0.3CH ₃ COOH+ 0.7O ₂ +0.7CH ₃ C(O)OOH	4.3×10 ⁻¹³ exp(1040/T) 13

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k135	APINO3+CH ₃ O ₂ → 0.335CH ₂ O+0.665CH ₂ O+ 0.665HO ₂ +0.665CH ₃ COCH ₃ + 1.33PROPEO2+CO ₂	4.52×10 ⁻¹²	13
k136	APINO3+CH ₃ CO ₃ → 2CO ₂ +CH ₃ O ₂ + CH ₃ COCH ₃ +2PROPEO2	4.6×10 ⁻¹² exp(530/T)	13
k137	APINO3+APINO3 → 2CO ₂ +2CH ₃ COCH ₃ +4PROPEO2	2.3×10 ⁻¹² exp(530/T)	13
k138	MACR+OH → 0.5MACRO2+0.5HO ₂ +0.5MCO3	1.86×10 ⁻¹¹ exp(175/T)	13
k139	MACR+O ₃ → 0.8CH ₃ COCHO+0.13HO ₂ + 0.37CO+0.1H ₂ +0.2OH+ 0.34CH ₂ O+0.14CO ₂	1.359×10 ⁻¹⁵ exp(-2112/T)	13
k140	MVK+OH → MACRO2	2.67×10 ⁻¹² exp(452/T)	13
k141	MVK+O3 → 0.05CH ₂ O+0.95CH ₃ COCHO+ 0.08OH+0.15HO ₂ +0.12H ₂ + 0.16CO ₂ +0.44CO	7.51×10 ⁻¹⁶ exp(-1521/T)	13
k142	MACRO2+NO → 0.015ONITR+0.985NO ₂ + 0.985HO ₂ +0.158CH ₂ O+ 0.158CH ₃ COCHO+0.828CO+ 0.828CH ₃ COCHO	2.7×10 ⁻¹² exp(360/T)	13
k143	MACRO2+NO ₃ → NO ₂ +HO ₂ +0.16CH ₂ O+ 0.16CH ₃ COCHO+0.84CO+ 0.84CH ₃ COCHO	2.4×10 ⁻¹²	13
k144	MACRO2+HO ₂ → MACROOH	1.9×10 ⁻¹³ exp(1300/T)	13
k145	MACRO2+CH ₃ O ₂ → 0.916HO ₂ +1.064CH ₂ O+ 0.458CO+0.458CH ₃ COCHO+ 0.229CH ₃ OH+0.458CH ₃ CHO+	4.1288×10 ⁻¹³	13
k146	MACRO2+CH ₃ CO ₃ → 0.794CO ₂ +0.794CH ₃ O ₂ + 0.412CH ₃ CHO+0.544CH ₂ O+ 0.794CH ₃ COCHO+0.794HO ₂ + 0.206CH ₃ COOH+0.25CO	2.475×10 ⁻¹²	13
k147	MACROOH+OH → MACRO2+H ₂ O	1.9×10 ⁻¹² exp(190/T)	13
k148	MACROOH+OH → 2CH ₃ CHO+OH+H ₂ O	3.9×10 ⁻¹⁷ exp(253/T)T ²	13
k149	MACROOH+OH → MCO3+H ₂ O	2.27×10 ⁻¹⁷ exp(696/T)T ²	13
k150	MACROOH+OH → 0.6C ₂ H ₅ OOH+HO ₂ + 0.4CH ₃ CHO+H ₂ O	5.16×10 ⁻¹⁷ exp(253/T)T ²	13
k151	MCO3+NO → CH ₃ CO ₃ +CH ₂ O+NO ₂	5.3×10 ⁻¹² exp(360/T)	13
k152	MCO3+NO ₂ +M → MPAN+M	k ₀ = 2.7×10 ⁻²⁸ (T/298) ^{-7.1} k _∞ = 1.2×10 ⁻¹¹ (T/298) ^{-0.9} F _c = 0.3	13

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k153	MPAN+M → MCO3+NO ₂ +M	$k_0 = 5.0 \times 10^{-2} \exp(-12875/T)$ $k_\infty = 2.2 \times 10^{+16} \exp(-13435/T)$ $F_c = 0.27$	13
k154	MCO3+NO ₃ → CH ₃ CO ₃ +CH ₂ O+NO ₂ +O ₂	5.0×10^{-12}	13
k155	MCO3+HO ₂ → 0.3O ₃ +0.3CH ₃ COOH+0.7O ₂ +0.7CH ₃ C(O)OOH	$4.3 \times 10^{-13} \exp(1040/T)$	13
k156	MCO3+CH ₃ O ₂ → 1.655CH ₂ O+0.665HO ₂ +0.665CH ₃ CO ₃ +0.665CO ₂	4.52×10^{-12}	13
k157	MCO3+CH ₃ CO ₃ → 2CO ₂ +CH ₃ O ₂ +CH ₂ O+CH ₃ CO ₃	$4.6 \times 10^{-12} \exp(530/T)$	13
k158	MCO3+MCO ₃ → 2CO ₂ +2CH ₃ O ₂ +2CH ₂ O+2CH ₃ CO ₃	$2.3 \times 10^{-12} \exp(530/T)$	13
k159	CH ₃ COCHO+OH → CH ₃ CO ₃ +CO+H ₂ O	$8.4 \times 10^{-13} \exp(830/T)$	17
k160	CH ₃ COCHO+NO ₃ → CH ₃ CO ₃ +CO+HNO ₃	$1.4 \times 10^{-12} \exp(-1860/T)$	13
k161	PCHO+OH → PCO3+H ₂ O	9.1×10^{-11}	13
k162	PCHO+NO ₃ → PCO3+HNO ₃	5.4×10^{-14}	13
k163	PCO3+NO → PROPEO2+NO ₂ +CO ₂ +XO2	$5.3 \times 10^{-12} \exp(360/T)$	13
k164	PCO3+NO ₂ +M → PCO3PAN+M	$k_0 = 2.7 \times 10^{-28} (T/298)^{-7.1}$ $k_\infty = 1.2 \times 10^{-11} (T/298)^{-0.9}$ $F_c = 0.3$	13
k165	PCO3PAN+M → PCO3+NO ₂ +M	$k_0 = 5.0 \times 10^{-2} \exp(-12875/T)$ $k_\infty = 2.2 \times 10^{+16} \exp(-13435/T)$ $F_c = 0.27$	13
k166	PCO3+NO ₃ → PROPEO2+NO ₂ +CO ₂ +XO2	5.0×10^{-12}	13
k167	PCO3+HO ₂ → 0.3O ₃ +0.3CH ₃ COOH+0.7O ₂ +0.7CH ₃ C(O)OOH	$4.3 \times 10^{-13} \exp(1040/T)$	13
k168	PCO3+CH ₃ O ₂ → 0.665PROPEO2+CH ₂ O+0.665HO ₂ +0.665XO2	4.52×10^{-12}	13
k169	PCO3+CH ₃ CO ₃ → 2CO ₂ +PROPEO2+XO2+CH ₃ O ₂	$4.6 \times 10^{-12} \exp(530/T)$	13
k170	PCO3+PCO ₃ → 2CO ₂ +2PROPEO2+2XO2	$2.3 \times 10^{-12} \exp(530/T)$	13
k171	ONITU+OH → 0.694ONITUO2+0.25HNO ₃ +0.25HO ₂ +0.3CH ₃ COCH ₃ +0.05NO ₂	1.83×10^{-12}	13
k172	ONITUO2+NO → 1.294NO ₂ +0.706ONITR+0.4HO ₂ +0.116CH ₂ O+0.386CH ₃ CHO+0.209MEK+0.395XO2	$2.7 \times 10^{-12} \exp(360/T)$	13
k173	ONITUO2+NO ₃ → 1.294NO ₂ +0.706ONITR+0.4HO ₂ +0.116CH ₂ O+0.386CH ₃ CHO+0.209MEK+0.395XO2	2.4×10^{-12}	13
k174	ONITUO2+HO ₂ → 0.7ONITR+0.3ONITUO2	$1.9 \times 10^{-13} \exp(1300/T)$	13
k175	ONITR+OH → MCO3+H ₂ O	1.5×10^{-11}	13

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k176	ONITR+NO ₃ → MCO3+HNO ₃	$1.4 \times 10^{-12} \exp(-1860/T)$	13
k177	MEK+OH → MEKO2	$3.24 \times 10^{-18} \exp(414/T)T^2$	13
k178	MEKO2+NO → NO ₂ +1.329CH ₃ CHO+0.6CH ₃ CO ₃ +0.07CH ₂ O+0.4HO ₂ +0.197MEK	$2.7 \times 10^{-12} \exp(360/T)$	13
k179	MEKO2+NO ₃ → NO ₂ +1.329CH ₃ CHO+0.6CH ₃ CO ₃ +0.07CH ₂ O+0.4HO ₂ +0.197MEK	2.4×10^{-12}	13
k180	MEKO2+HO ₂ → MEKOOH	$1.9 \times 10^{-13} \exp(1300/T)$	13
k181	MEKO2+CH ₃ O ₂ → 0.305CH ₃ OH+0.699HO ₂ +0.75CH ₂ O+0.08CH ₃ CO ₃ +0.295MEK+0.654CH ₃ CHO+0.042CH ₃ COCHO	9.764×10^{-13}	13
k182	MEKOOH+OH → MEKO2+H ₂ O	$1.9 \times 10^{-12} \exp(190/T)$	13
k183	MEKOOH+OH → MEK+ OH+H ₂ O	$1.17 \times 10^{-17} \exp(696/T)T^2$	13
k184	MEKOOH+OH → CH ₃ CHO+0.5MEK+OH+H ₂ O	$9.75 \times 10^{-17} \exp(253/T)T^2$	13
k185	MEKOOH+OH → CH ₃ COCHO+OH+H ₂ O	$3.28 \times 10^{-18} \exp(253/T)T^2$	13
k186	ALKEN+OH → ALKENO2	$9.19 \times 10^{-12} \exp(-522.22/T)$	13
k187	ALKEN+O ₃ → 0.9CH ₃ CHO+0.23ALKENO2+0.09CH ₃ COCH ₃ +0.34CH ₃ O ₂ +0.08CH ₄ +0.02C ₂ H ₆ +0.3CO+0.01CH ₃ OH+0.42OH	$4.95 \times 10^{-15} \exp(-1054.84/T)$	13
k188	ALKEN+NO ₃ → ALKENO2+NO ₂	$3.95 \times 10^{-12} \exp(-327.93/T)$	13
k189	ALKENO2+NO → 0.034ONITU+0.406CH ₂ O+1.666CH ₃ CHO+0.966NO ₂ +0.38CH ₃ COCH ₃ +0.966HO ₂	$2.7 \times 10^{-12} \exp(360/T)$	13
k190	ALKENO2+NO ₃ → 0.393CH ₃ COCH ₃ +NO ₂ +HO ₂ +0.724CH ₃ CHO+0.42CH ₂ O	2.4×10^{-12}	13
k191	ALKENO2+HO ₂ → ALKENOOH	$1.9 \times 10^{-13} \exp(1300/T)$	13
k192	ALKENO2+CH ₃ O ₂ → 0.305CH ₃ OH+0.265CH ₃ CHO+0.695CH ₂ O+0.06CH ₃ COCH ₃ +0.305CH ₃ COCHO+0.78HO ₂	1.22×10^{-13}	13
k193	ALKENOOH+OH → ALKENO2+H ₂ O	$1.9 \times 10^{-12} \exp(190/T)$	13
k194	ALKENOOH+OH → CH ₃ COCHO+OH+H ₂ O	$9.46 \times 10^{-17} \exp(253/T)T^2$	13
k195	ALKAN+OH → ALKANO2	$1.63 \times 10^{-17} \exp(385.22/T)T^2$	13
k196	ALKANO2+NO → 0.007CH ₂ O+0.362CH ₃ CHO+0.289CH ₃ COCH ₃ +0.799NO ₂ +0.412MEK+0.082CH ₃ O ₂ +0.2ONITU+0.268XO2+ 0.449HO ₂	$2.7 \times 10^{-12} \exp(260/T)$	13

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Table 4. Continued.

#	reaction	rate coefficient	refs.
k197	ALKANO ₂ +NO ₃ → NO ₂ +0.562HO ₂ +0.336XO ₂ + 0.101CH ₃ O ₂ +0.517MEK+ 0.001CH ₂ O+0.454CH ₃ CHO	2.4×10 ⁻¹²	13
k198	ALKANO ₂ +HO ₂ → ALKANO ₂ OOH	1.9×10 ⁻¹³ exp(1300/T)	13
k199	ALKANO ₂ +CH ₃ O ₂ → 0.045CH ₃ COCH ₃ +0.626HO ₂ + 0.305CH ₃ OH+0.696CH ₂ O+ 0.315MEK+0.012CH ₃ O ₂ + 0.442CH ₃ CHO+0.14XO ₂	3.7652×10 ⁻¹³	13
k200	ALKANO ₂ OOH+OH → ALKANO ₂ +H ₂ O	1.9×10 ⁻¹² exp(190/T)	13
k201	ALKANO ₂ OOH+OH → CH ₃ CHO+OH+H ₂ O	1.07×10 ⁻¹⁷ exp(253/T)	13
k202	ALKANO ₂ OOH+OH → MEK+OH+H ₂ O	3.82×10 ⁻¹⁷ exp(696/T)T ²	13
k203	AROM+OH → 0.77AROMO ₂ +0.212HO ₂	1.01×10 ⁻¹¹ exp(58.45/T)	13
k204	AROMO ₂ +NO → 0.423CH ₃ COCHO+NO ₂ + 0.658CH ₃ CO ₃ +0.658CO+1.658HO ₂	2.7×10 ⁻¹² exp(360/T)	13
k205	AROMO ₂ +NO ₃ → 0.423CH ₃ COCHO+NO ₂ + 0.658CH ₃ CO ₃ +0.658CO+1.658HO ₂	2.4×10 ⁻¹²	13
k206	AROMO ₂ +HO ₂ → AROMOOH	1.9×10 ⁻¹³ exp(1300/T)	13
k207	AROMO ₂ +CH ₃ O ₂ → 0.087*CH ₃ COCHO + 0.135*CO+ 0.135*CH ₃ CO ₃ + 0.305*CH ₃ OH+ 0.695*CH ₂ O + 0.915*HO ₂	2.31×10 ⁻¹³	13
k208	AROMO ₂ OOH+OH → AROMO ₂ +H ₂ O	1.9×10 ⁻¹² exp(190/T)	13
k209	AROMO ₂ OOH+OH → OH+H ₂ O	4.61×10 ⁻¹⁸ exp(253/T)T ²	13
k210	AROMO ₂ OOH+OH → CH ₃ CO ₃ +CO+OH+HO ₂ +H ₂ O	4.19×10 ⁻¹⁷ exp(696/T)T ²	13
k211	XO ₂ +NO → NO ₂ +HO ₂	2.7×10 ⁻¹² exp(360/T)	13
k212	XO ₂ +NO ₃ → NO ₂ +HO ₂	2.4×10 ⁻¹²	13
k213	XO ₂ +HO ₂ → XOOH	1.9×10 ⁻¹³ exp(1300/T)	13
k214	XO ₂ +CH ₃ O ₂ → 0.305CH ₃ O ₂ +0.695CH ₂ O+0.39HO ₂	1.22×10 ⁻¹³	13
k215	XO ₂ +CH ₃ CO ₃ → 0.275CH ₃ COOH+0.725CO ₂ + 0.725CH ₃ O ₂	7.37×10 ⁻¹³	13
k216	XOOH+OH → XO ₂ +H ₂ O	1.9×10 ⁻¹² exp(190/T)	13
k217	XOOH+OH → OH+H ₂ O	7.69×10 ⁻¹⁷ exp(253/T)T ²	13

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Table 4. Continued.

T = temperature (K); M = air density (molecules cm⁻³)

Rate coefficients are in cm³ molecules⁻¹ s⁻¹ for bimolecular reactions and in cm⁶ molecules⁻² s⁻¹ for termolecular reactions.

In the latter case, the rate coefficient is defined by

$$k(T, M) = \left[\frac{k_0(T) \cdot [M]}{1 + \frac{k_0(T) \cdot [M]}{k_{\infty}(T)}} \right] \cdot F_c \left\{ 1 + \left[\log_{10} \left(\frac{k_0(T) \cdot [M]}{k_{\infty}(T)} \right) \right]^2 \right\}^{-1}$$

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- 1, Atkinson et al. (1997); 2, DeMore et al. (1997); 3, Jenkin and Cox (1987); 4, Matzkies and Manthe (1998)
- 5, Yu and Varandas (1997); 6, Hall et al. (1988); 7, Matsumi et al. (1993); 8, Greenberg and Heicklen (1972)
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- 13, Brocheton (1999); 14, Brasseur et al. (1998); 15, Hauglustaine et al. (1998); 16, McCabe et al. (2001)
- 17, Tyndall et al. (1995)

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Table 5. Global surface emissions of trace gases in LMDz-INCA.

Species	Emission	Species	Emission
NO_x (TgN yr⁻¹)		C₄- and higher alkanes (TgC yr⁻¹)	
Fossil Fuel + Industry + Indst.	22.87	Fossil Fuel + Industry + Indst.	19.99
Biofuel		Biofuel	
Biomass Burning	9.20	Biomass Burning	1.97
Soils	5.48	<i>Total</i>	21.96
Lightning	5.00	Aromatics (TgC yr⁻¹)	
International Shipping	3.08	Fossil Fuel + Industry + Indst.	7.82
		Biofuel	
Aircraft	0.70	Biomass Burning	4.69
<i>Total</i>	46.33	<i>Total</i>	12.51
N₂O (TgN yr⁻¹)		Ethene (TgC yr⁻¹)	
Soils	7.20	Biomass Burning	8.63
Ocean	3.50	Fossil Fuel + Industry + Indst.	4.45
		Biofuel	
Execra	1.02	Vegetation	4.30
Biomass Burning	0.83	Ocean	0.67
Fossil Fuel + Industry + Indst.	0.73	<i>Total</i>	18.05
Biofuel		Propene (TgC yr⁻¹)	
Deforestation	0.36	Biomass Burning	3.14
<i>Total</i>	13.64	Fossil Fuel + Industry + Indst.	1.68
CO (Tg yr⁻¹)		Biofuel	
Biomass Burning	994.56	Vegetation	0.80
Fossil Fuel + Industry + Indst.	488.46	Ocean	0.34
Biofuel		<i>Total</i>	5.96
Ocean	50.00	C₄- and higher alkenes (TgC yr⁻¹)	
<i>Total</i>	1533.03	Biomass Burning	6.00
CH₄ (Tg yr⁻¹)		Fossil Fuel + Industry + Indst.	2.40
Wetlands, Swamps, and Bogs	177.78	Biofuel	
		Vegetation	1.70
Fossil Fuel + Industry + Indst.	94.72	Ocean	0.28
Biofuel		<i>Total</i>	10.38
Ruminants	93.21	Ethyne (TgC yr⁻¹)	
Rice Cultivation	79.56	Biomass Burning	2.67
Biomass Burning	36.36	Fossil Fuel + Industry + Indst.	1.02
Landfills	35.70	Biofuel	
Termites	20.00	Vegetation	0.80
Tundra	3.24	Ocean	0.20
<i>Total</i>	540.57		

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Table 5. Continued.

Species	Emission	Species	Emission
Isoprene (TgC yr ⁻¹)		<i>Total</i>	4.69
Vegetation	410.67	Acetone (TgC yr ⁻¹)	
Ocean	0.88	Vegetation	34.72
<i>Total</i>	411.55	Ocean	12.41
Terpenes (TgC yr ⁻¹)		Biomass Burning	2.00
Vegetation	96.06	Fossil Fuel + Industry + Indst.	0.63
<i>Total</i>	96.06	Biofuel	
Methanol (TgC yr ⁻¹)		<i>Total</i>	49.76
Vegetation	86.94	Formaldehyde (TgC yr ⁻¹)	
Plant Decay	13.00	Biomass Burning	2.39
Biomass Burning	4.30	Fossil Fuel + Industry + Indst.	0.25
<i>Total</i>	104.24	Biofuel	
Ethane (TgC yr ⁻¹)		<i>Total</i>	2.64
Biomass Burning	4.61	C ₃ - and higher aldehyde (TgC yr ⁻¹)	
Fossil Fuel + Industry + Indst.	3.18	Biomass Burning	3.40
Biofuel		Fossil Fuel + Industry + Indst.	0.35
Vegetation	0.80	Biofuel	
<i>Total</i>	8.59	<i>Total</i>	3.75
Propane (TgC yr ⁻¹)		Ketones other than acetone (TgC yr ⁻¹)	
Fossil Fuel + Industry + Indst.	6.95	Fossil Fuel + Industry + Indst.	0.95
Biofuel		Biofuel	
Vegetation	1.60	<i>Total</i>	0.95
Biomass Burning	1.26	Hydrogen (Tg yr ⁻¹)	
<i>Total</i>	9.81	Fossil Fuel + Industry + Indst.	16.00
		Biofuel	
		Biomass Burning	13.00
		Ocean and Soils (50% each)	10.00

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Table 6. Break-down of annual average hydroxyl radical concentrations classified by tropospheric subdomains (10^5 molecules cm^{-3}). Horizontal domains: northern extratropics (NET, 90° – 30° N), tropics (TRO, 30° N– 30° S), southern extratropics (SET, 30° – 90° S); vertical domains: planetary boundary layer (PBL, below 750 hPa), free troposphere (FT, 750–500 hPa), upper troposphere and tropopause region (UT, 500–250 hPa), respectively. The rows and columns denoted by *TOT* and **TOT**, respectively, refer to the mean calculated over the specific horizontal layer or altitude range.

subdomain	PBL	FT	UT	<i>TOT</i>
NET	9.0	6.4	5.6	7.8
TRO	16.6	10.7	7.7	12.5
SET	5.3	4.4	4.5	5.4
TOT	11.9	8.1	6.4	9.6

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Table 7. Global Tropospheric Ozone Budget in LMDz-INCA ($\text{Tg O}_3 \text{ yr}^{-1}$).

	Global	Northern Hemisphere	Southern Hemisphere
Photochemical production	4436	2773	1663
Stratospheric influx	715	366	349
Photochemical destruction	−3890	−2304	−1586
Dry deposition	−1261	−835	−426
Net photochemistry	546	469	77
Ozone Burden (Tg O_3)	303	178	125

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Table 8. Carbon monoxide budget for the global troposphere defined as the region extending from the surface up to the 250 hPa altitude level. LMDz-INCA results are compared to the a posteriori estimates obtained by an inverse modelling study of the global CO cycle as given by Bergamaschi et al. (2000). All numbers are written as TgCO yr⁻¹. The average and range (in brackets) of estimates for the study of Bergamaschi et al. is denoted. Percent values in parentheses indicate VOC-to-CO conversion efficiencies calculated with the LMDz-INCA model.

	LMDz-INCA	Bergamaschi et al. (2000)
Global Source	3019	2818 (2244–3392)
<i>Primary Source</i>	1533	1349 (1022–1675)
Biomass Burning	811	674 (506–840)
Technological	672	656 (511–800)
Oceanic	50	19 (5–35)
<i>RCC^a Oxidation</i>	1489 (55%)	1468 (1220–1715)
Methane ^b	854 (90%)	831
Isoprene	359 (37%)	348 (231–464)
Methanol	110 (45%)	–
Terpenes	49 (22%)	145 (98–191)
Acetone	28 (24%)	–
∑ other VOC ^c	89 (38%)	145 (61–229)
Global Sink	3019	2938 (2863–3010)
Photochemical		
Destruction	2653	2471 (2415–2529)
Dry Deposition	135	301 (292–308)
Stratospheric		
Entrainment	231	166 (156–173)

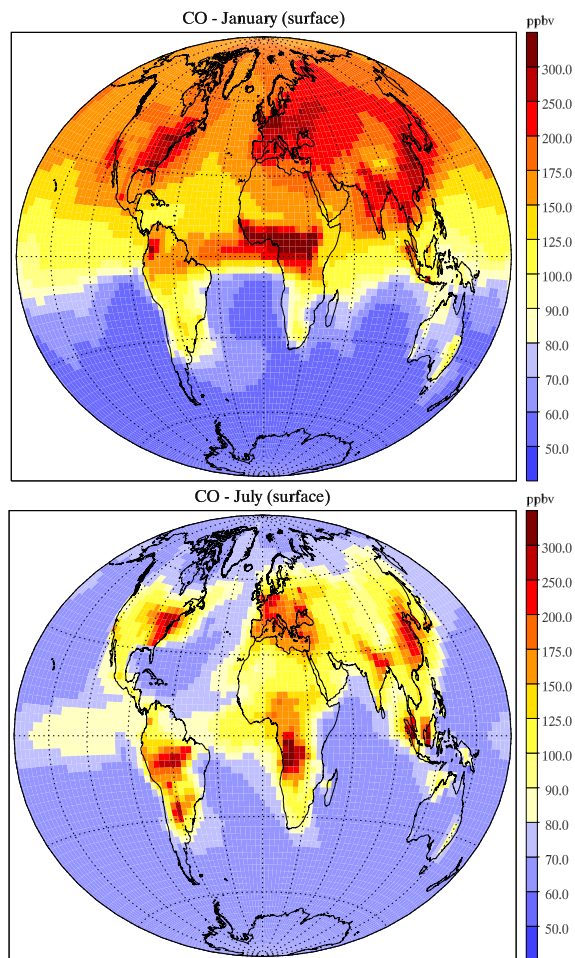
^a reactive carbon compounds (cf. text for definition)

^b fixed in Bergamaschi et al. (2000); calculated in LMDz-INCA

^c mostly anthropogenic

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**Fig. 1.** Carbon monoxide surface mixing ratio for January and July (ppbv).[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Print Version](#)[Interactive Discussion](#)

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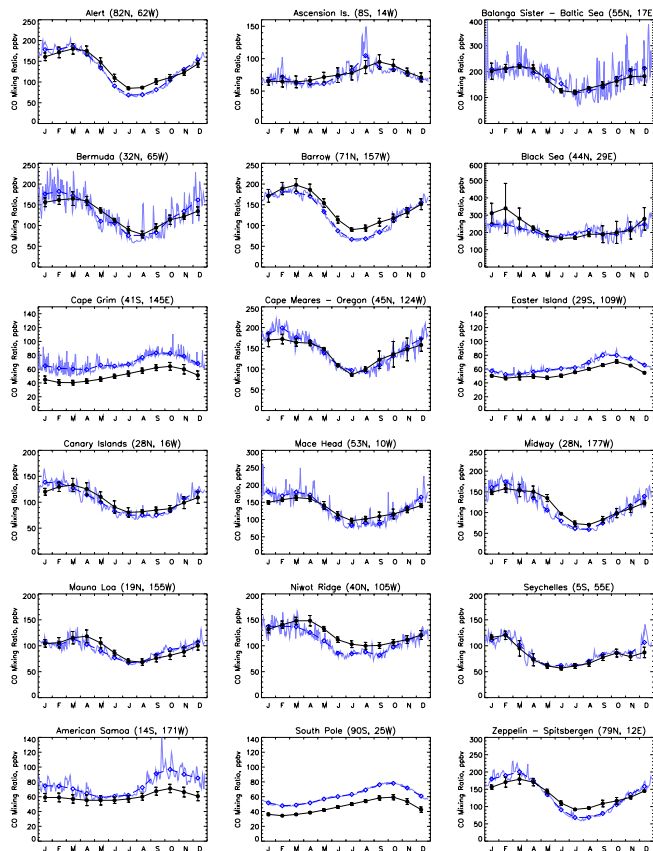


Fig. 2. Comparison of observed and calculated monthly mean CO mixing ratios at the surface. Black circles and solid lines denote observed values from [Novelli et al. \(2003\)](#) averaged over 5 to 10 years depending on the sites; vertical bars are standard deviations. Blue open diamonds and heavy dashed lines represent monthly mean CO concentrations calculated by the model. Solid thin blue lines depict the daily variation in the model data.

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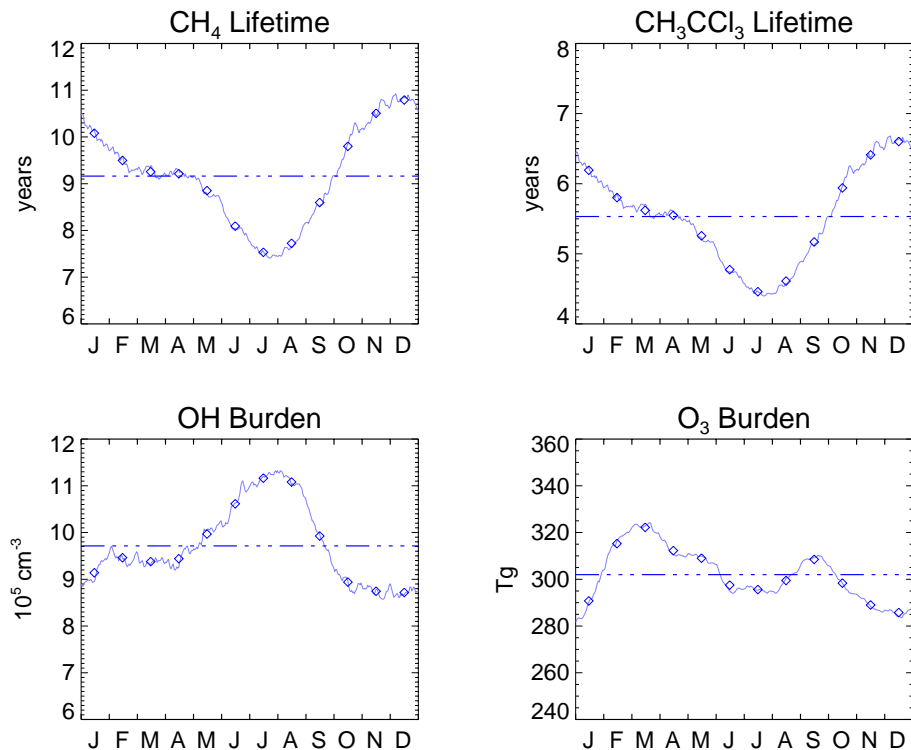


Fig. 3. Seasonal cycle of the globally averaged CH₄ and CH₃CCl₃ tropospheric lifetimes, OH abundance, and O₃ burden for the entire troposphere. Also shown are annual mean values of the above quantities (dashed-dotted lines).

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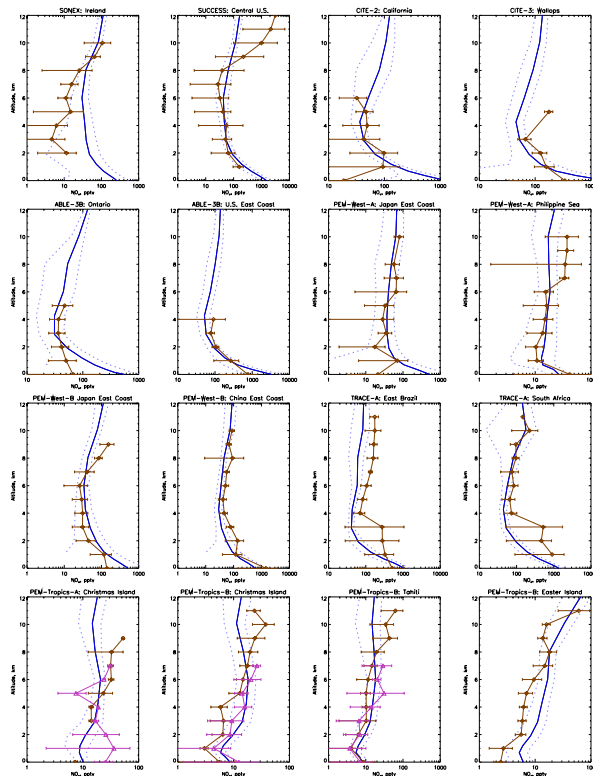


Fig. 4. Comparison of calculated and observed (Emmons et al., 2000) vertical profiles of the NO_x mixing ratio (pptv). Model results are shown as solid blue lines with dotted blue lines marking one standard deviation. Open diamonds represent mean observed values (with horizontal bars denoting the standard deviation). In case an additional data set is available deriving from a second aircraft used during a specific campaign, these data have been included and are denoted by open triangles. Simulated values have been sampled from the model output over the same region and month as the observational data.

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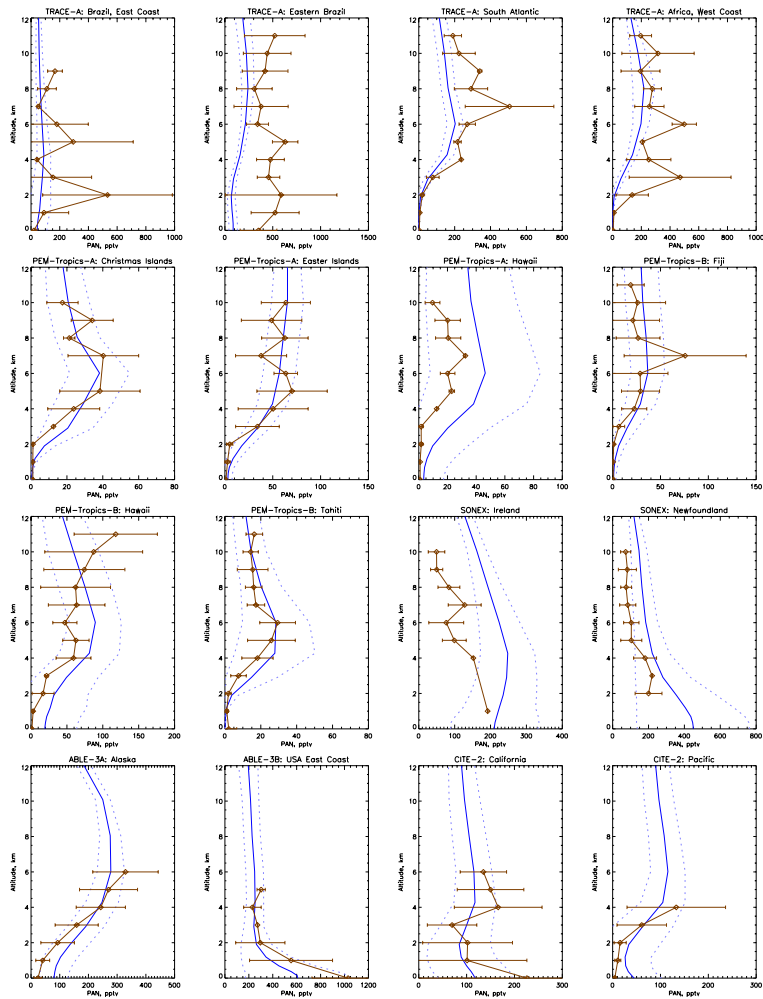
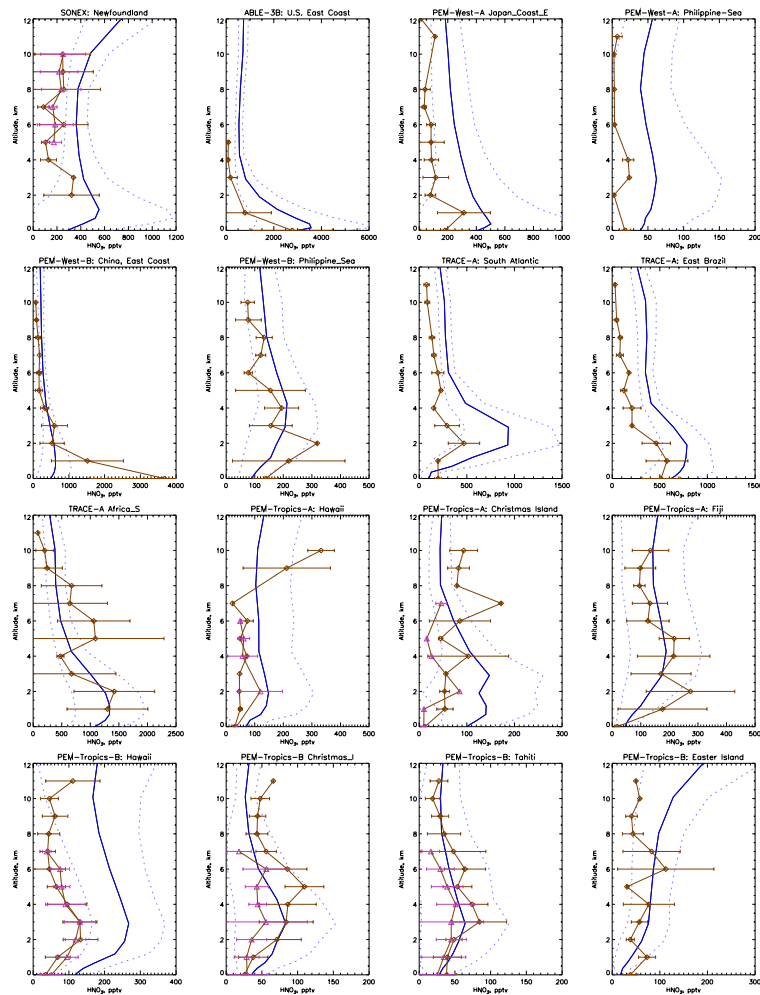


Fig. 5. Same as Fig. 4, but for PAN.

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Fig. 6. Same as Fig. 4, but for HNO_3 .

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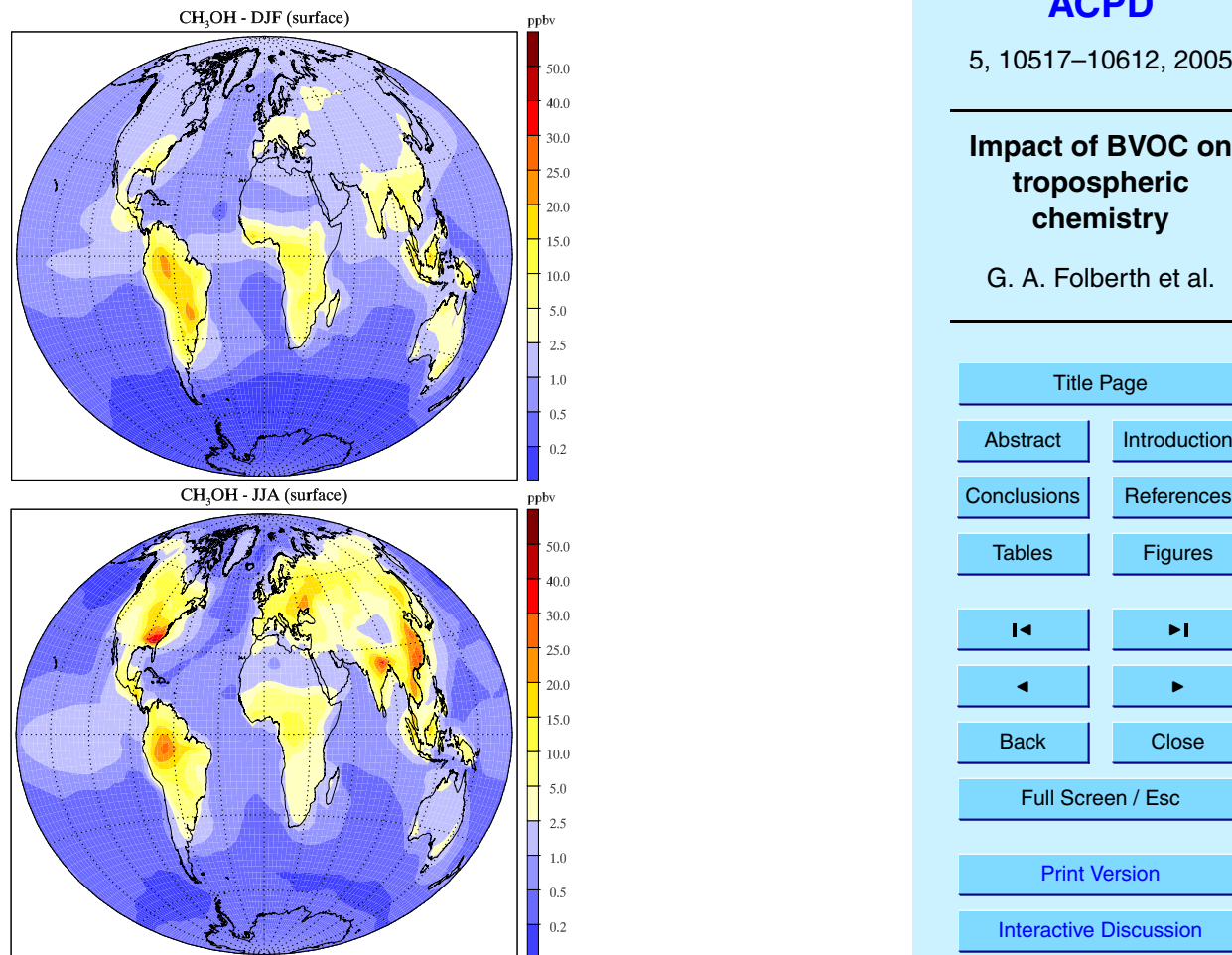
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**Fig. 7.** Methanol surface mixing ratio for winter and summer season (ppbv).[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Print Version](#)[Interactive Discussion](#)

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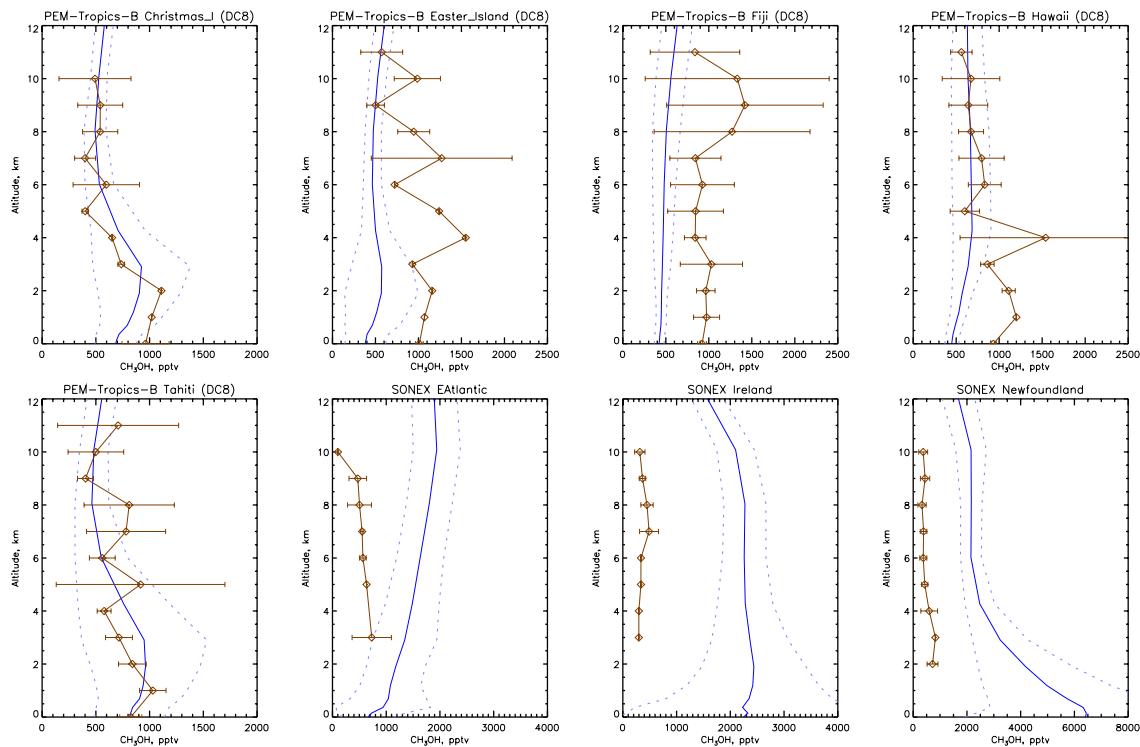


Fig. 8. Same as Fig. 4, but for methanol.

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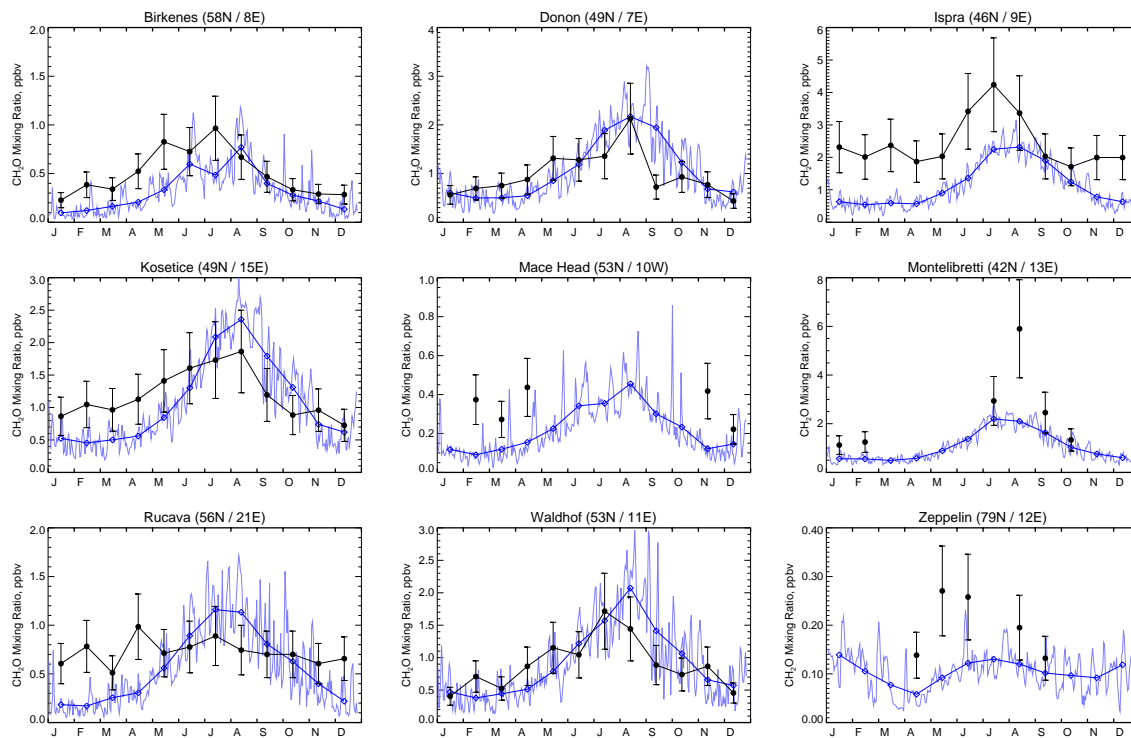


Fig. 9. Same as Fig. 2, but for CH₂O.

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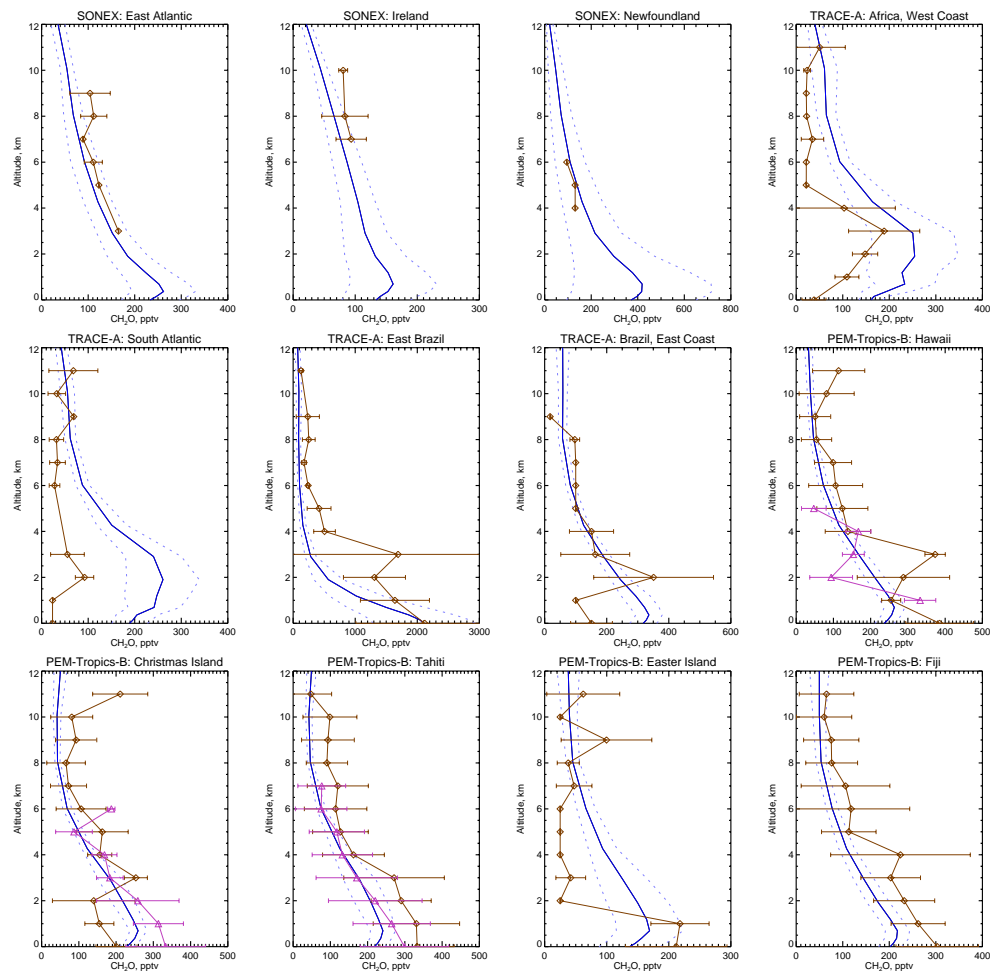


Fig. 10. Same as Fig. 4, but for formaldehyde.

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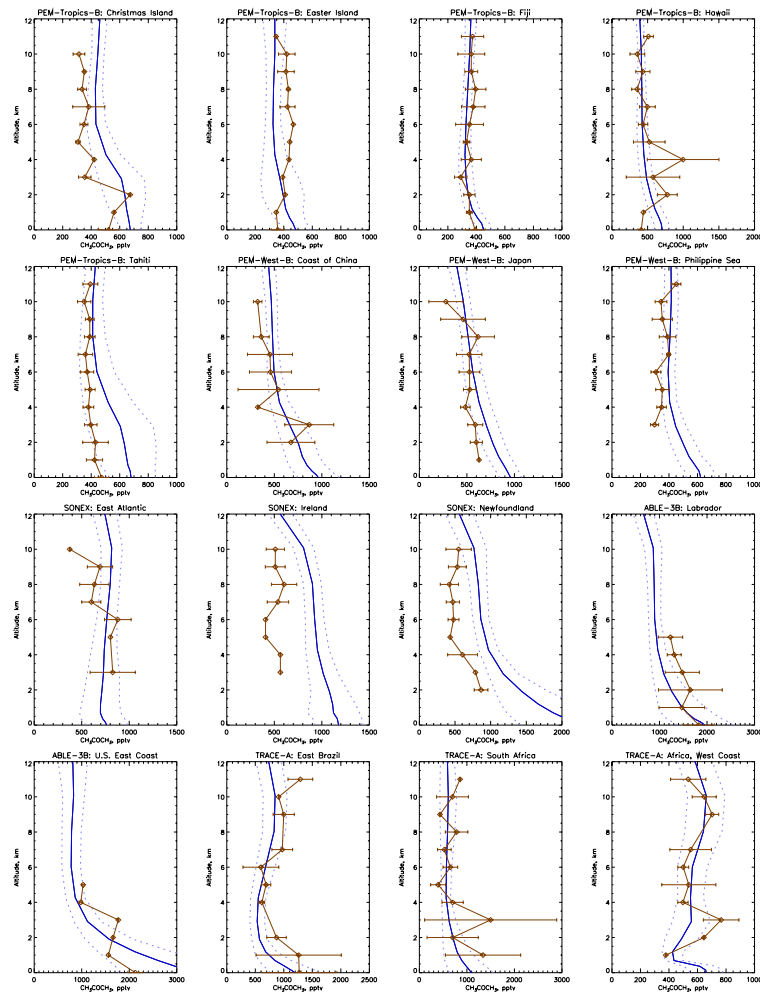


Fig. 11. Same as Fig. 4, but for acetone.

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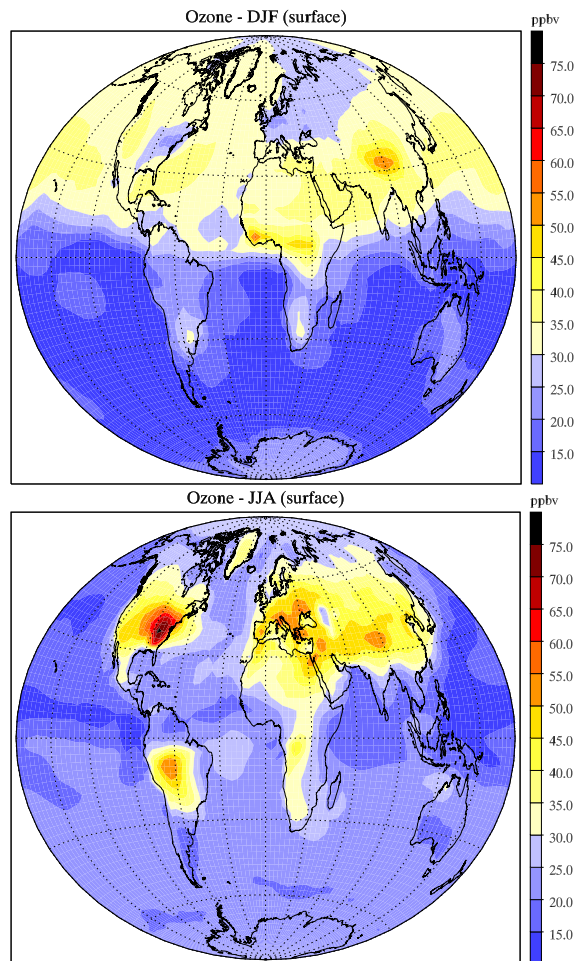
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**Fig. 12.** Ozone surface mixing ratio for winter and summer season (ppbv).[Title Page](#)[Abstract](#)[Introduction](#)[Conclusions](#)[References](#)[Tables](#)[Figures](#)[◀](#)[▶](#)[◀](#)[▶](#)[Back](#)[Close](#)[Full Screen / Esc](#)[Print Version](#)[Interactive Discussion](#)

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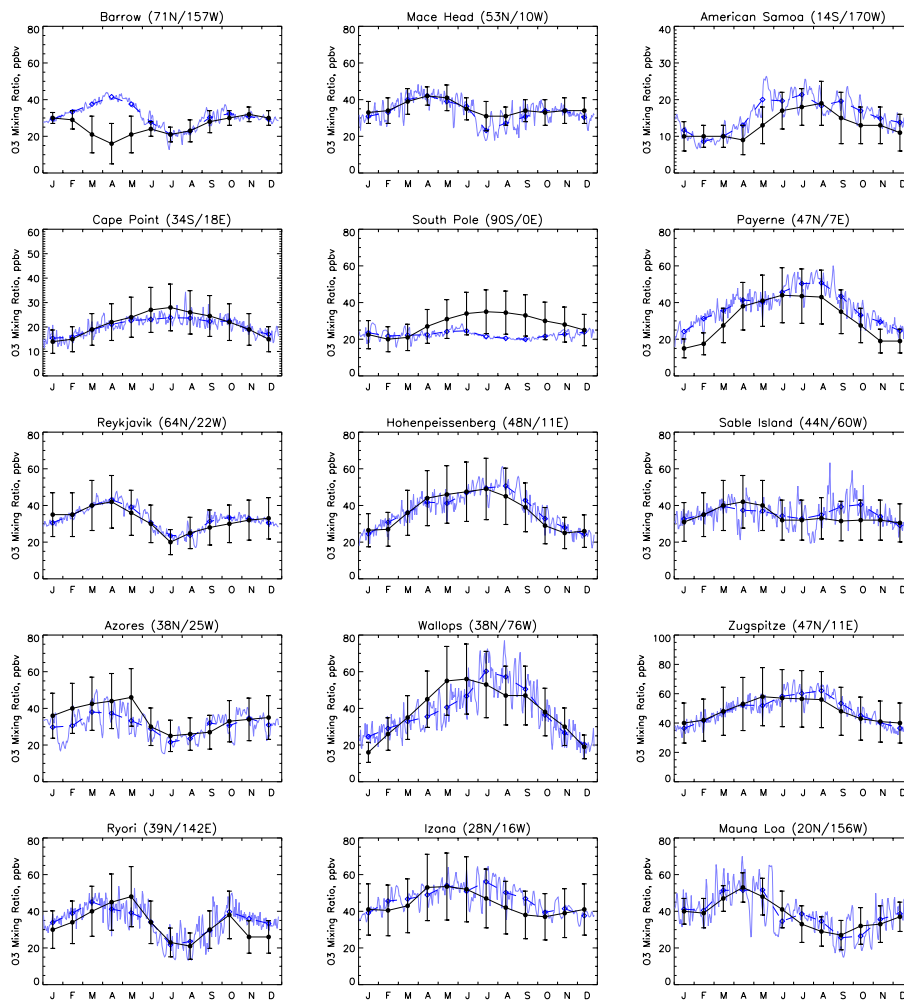
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Fig. 13. Same as Fig. 2, but for O₃.

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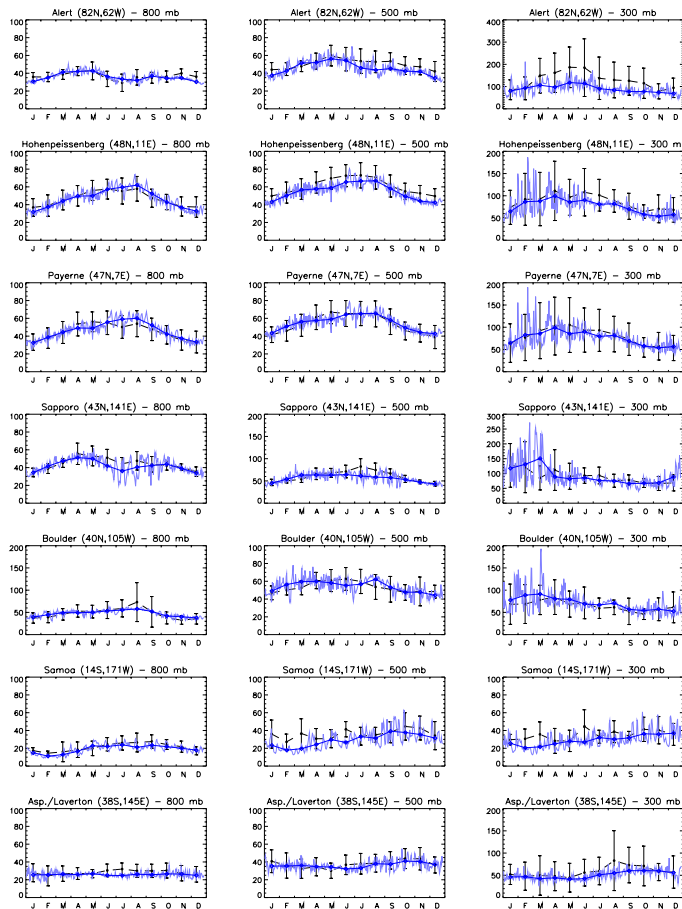


Fig. 14. Measured (solid circles) and calculated (solid light blue lines indicate day-to-day variations, open diamonds denote monthly mean values) seasonal cycle of O_3 mixing ratios (ppbv) at 7 stations and at the 800, 500, and 300 hPa levels. Observations were compiled by Logan (1999).

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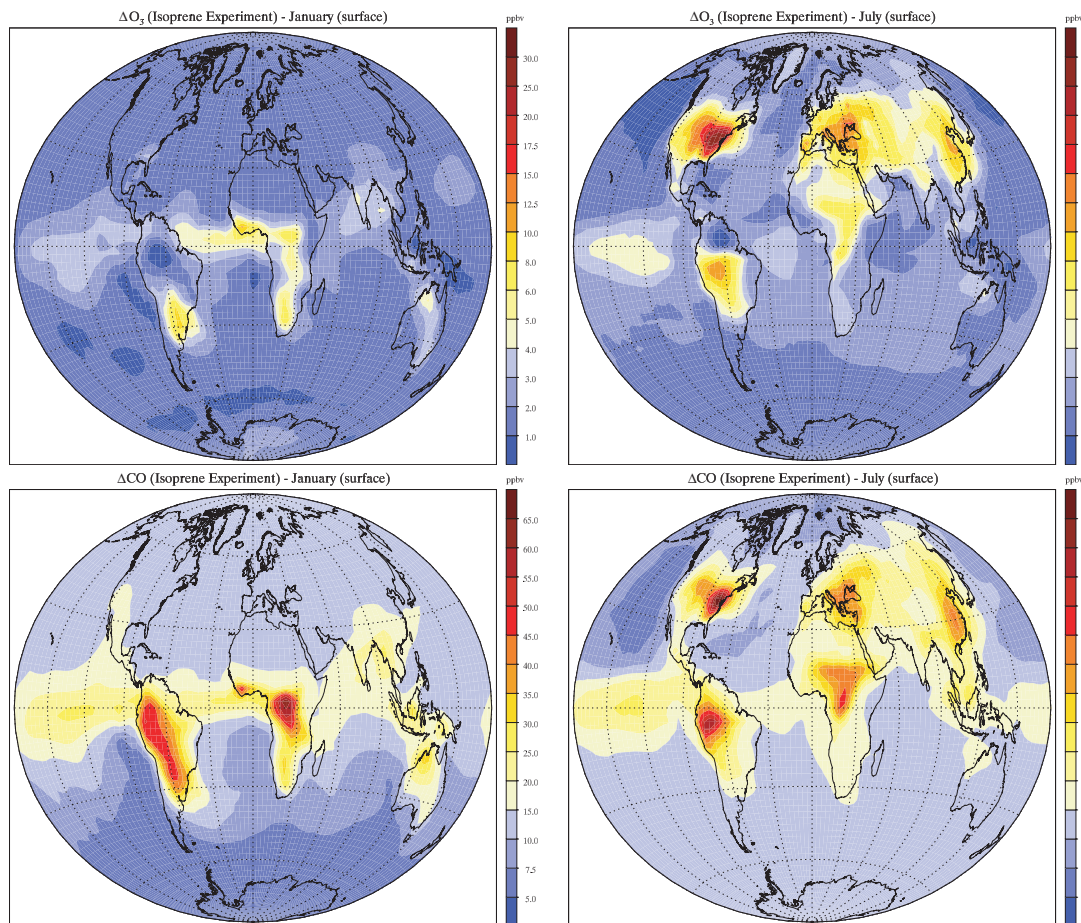


Fig. 15. Change in ozone (ΔO_3) and CO (ΔCO) surface mixing ratios (ppbv) due to isoprene for January and July.

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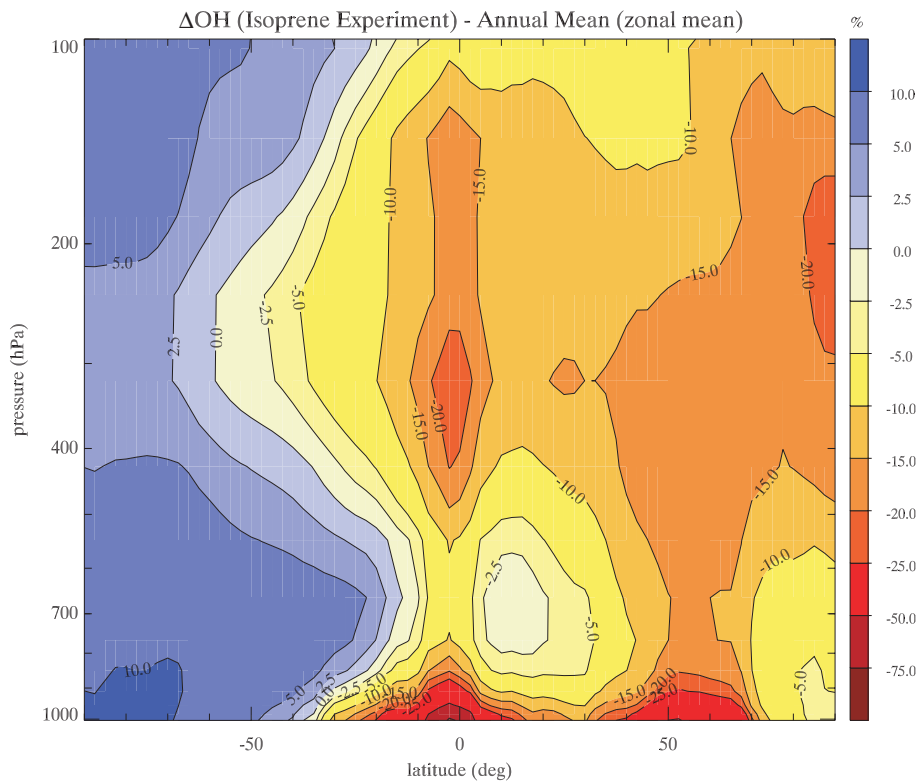


Fig. 16. Change in the zonally averaged annual mean OH distribution (ΔOH) due to isoprene (%).

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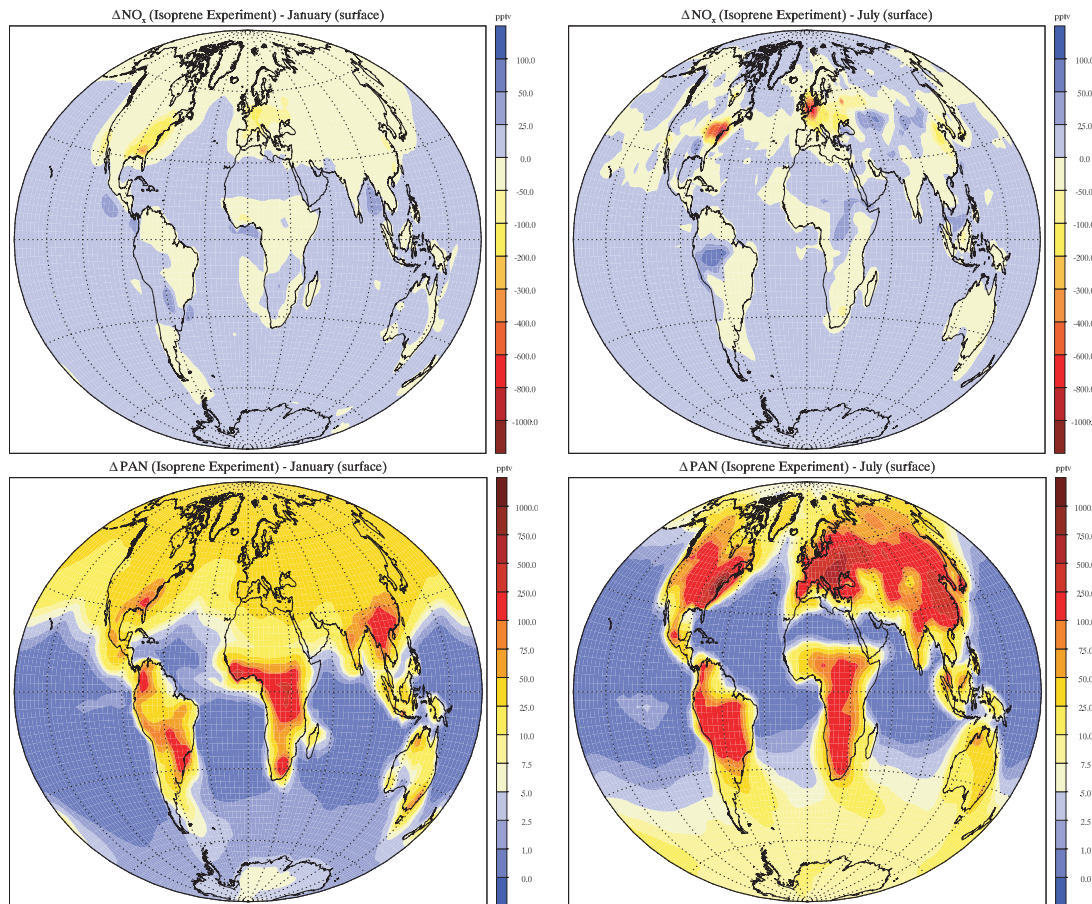


Fig. 17. Change in NO_x (ΔNO_x) and PAN (ΔPAN) surface mixing ratios (pptv) due to isoprene for January and July.

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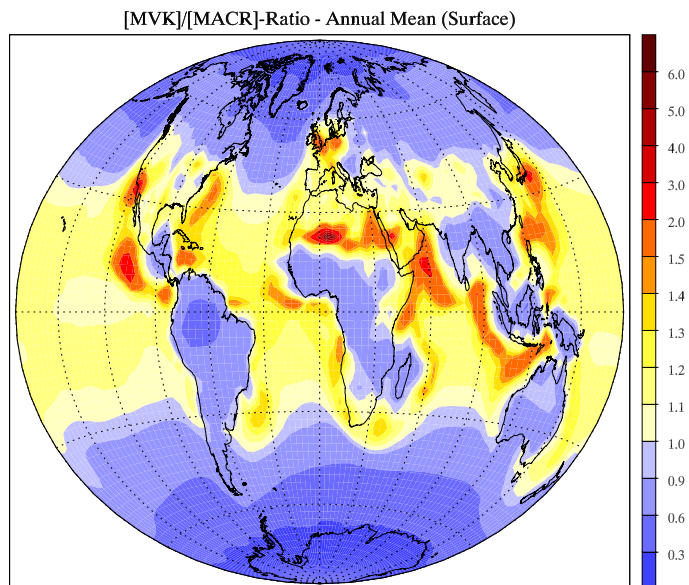


Fig. 18. Ratio of the annual mean methyl vinyl ketone and methacrolein surface mixing ratios, [MVK]/[MACR].

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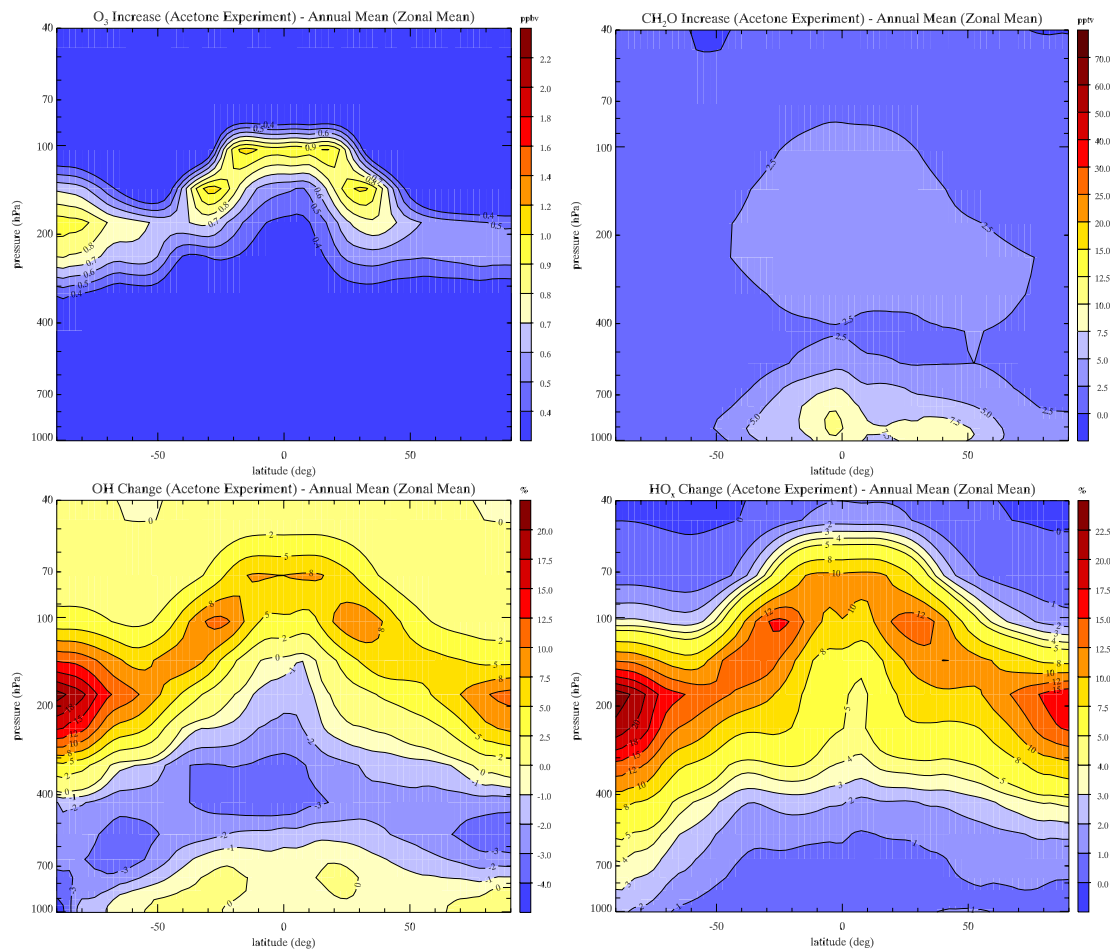


Fig. 19. Calculated annually and zonally averaged changes in O₃ (ppbv), CH₂O (ppbv), OH (%), and HO_x (%) due to acetone biogenic surface emissions (47 TgC yr⁻¹).

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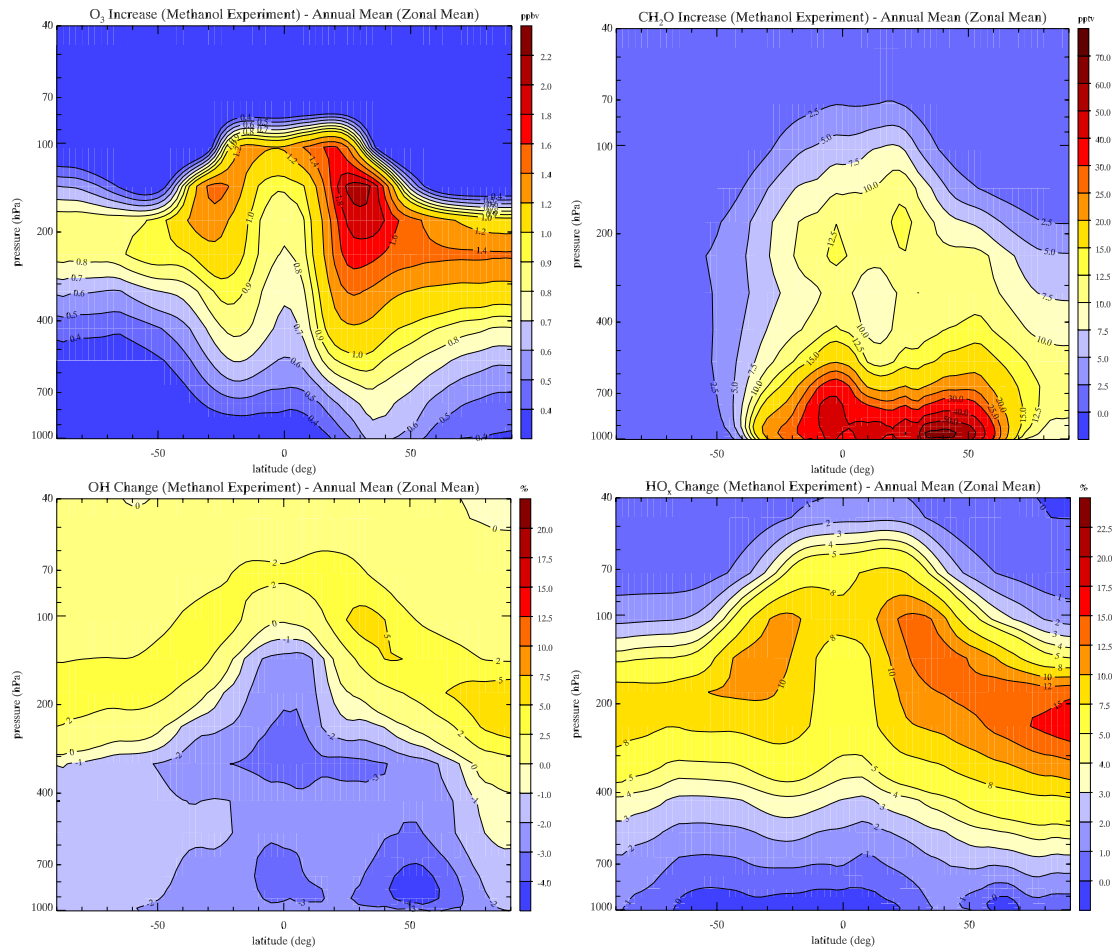


Fig. 20. Calculated annually and zonally averaged changes in O₃ (ppbv), CH₂O (ppbv), OH (%), and HO_x (%) due to methanol biogenic surface emissions (100 TgC yr⁻¹).

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