

Impact of primary formaldehyde on air pollution in the Mexico City Metropolitan Area

W. Lei^{1,2}, M. Zavala^{1,2}, B. de Foy^{1,3}, R. Volkamer^{2,4}, M. J. Molina^{2,5}, and L. T. Molina^{1,2}

¹Molina Center for Energy and the Environment, CA, USA

²Department of Earth, Atmospheric and Planetary Sciences, Massachusetts Institute of Technology, MA, USA

³Department of Earth and Atmospheric Sciences, Saint Louis University, MO, USA

⁴Department of Chemistry and Biochemistry, University of Colorado at Boulder, CO, USA

⁵Department of Chemistry and Biochemistry, University of California, San Diego, La Jolla, CA, USA

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Abstract. Formaldehyde (HCHO) is a radical source that plays an important role in urban atmospheric chemistry and ozone formation. The Mexico City Metropolitan Area (MCMA) is characterized by high anthropogenic emissions of HCHO (primary HCHO), which together with photochemical production of HCHO from hydrocarbon oxidation (secondary HCHO), lead to high ambient HCHO levels. The CAMx chemical transport model was employed to evaluate the impact of primary HCHO on its ambient concentration, on the RO_x radical budget, and on ozone (O₃) formation in the MCMA. Important radical sources, including HCHO, HONO, and O₃-olefin reactions, were constrained by measurements from routine observations of the local ambient air monitoring network and the MCMA-2003 field campaign. Primary HCHO was found not only to contribute significantly to the ambient HCHO concentration, but also to enhance the radical budget and O₃ production in the urban atmosphere of the MCMA. Overall in the urban area, total daytime radical production is enhanced by up to 10% and peak O₃ concentration by up to 8%; moreover primary HCHO tends to make O₃ both production rates and ambient concentration peak half an hour earlier. While primary HCHO contributes predominantly to the ambient HCHO concentration between nighttime and morning rush hours, significant influence on the radical budget and O₃ production starts early in the morning, peaks at mid-morning and is sustained until early afternoon.

1 Introduction

Formaldehyde (HCHO) is the most abundant carbonyl compound in both urban areas and the remote troposphere (Lurmann and Main, 1992; Ho et al., 2002; Hellen et al., 2004). It can be directly emitted from incomplete combustion processes (primary HCHO), such as combustion engines and biomass burning, or produced by photooxidation of hydrocarbons (secondary HCHO). Small amounts of HCHO can also be emitted directly from vegetation (Kesselmeier et al., 1997). The major loss processes of HCHO are photolysis and reaction with the hydroxyl radical (OH). HCHO serves as an important primary source for the hydroperoxy radical (HO₂) and contributes to other odd hydrogen radicals (RO_x=OH+HO₂+RO₂) through the radical propagation processes, which leads to ozone (O₃) production in the presence of nitrogen oxides (NO_x). Because of its widespread presence and its role in O₃ and radical photochemistry, HCHO is an essential component for assessing photochemical processes and radical budgets on local, regional and global scales. Unlike most primary volatile organic compounds (VOCs), photolysis of primary HCHO does not consume radicals, which makes primary HCHO a unique radical source. HCHO is a known carcinogen (Seinfeld and Pandis, 1998; Molina and Molina, 2002; IARC, 2004); exposure to high levels of ambient HCHO is detrimental to human health and to the ecosystem.

High concentrations of HCHO have been measured in the Mexico City Metropolitan Area (MCMA) (Molina and Molina, 2002). Grutter et al. (2005) and Volkamer et al. (2005) report that monthly average concentrations range



Correspondence to: W. Lei
(wlei@mit.edu)

from 4–9 ppb with monthly maxima up to 35 ppb in recent years, which are comparable or higher than those in many US urban cities (EPA, 2005). For example, the annual average concentrations of HCHO were 2.0 ppb in Chicago, 7.2 ppb in Los Angeles, 4.4 ppb in St. Louis, and 7.9 ppb in Houston in 2005. The high ambient level of HCHO in the MCMA is partially attributed to the direct emissions from mobile and industrial sources (Garcia et al. 2006). According to the emission estimates constructed from the official emissions inventory for the year 2002 in the MCMA (CAM, 2004), mobile and industrial sources are responsible for total HCHO emissions of about 2050 tons/yr in the MCMA, of which mobile emissions account for about 50%. There is evidence that the mobile emission rates of HCHO in the MCMA and the HCHO/CO₂ emission ratio are much higher than in the US urban areas (Kolb et al., 2004; Zavala et al., 2006). This is probably due to the differences in vehicle fleet characteristics (e.g., fleet age and emission control technology) (Zavala et al., 2006) and the reduced combustion efficiency attributed to high altitude (2.2 km a.s.l. in Mexico City). The high emissions of HCHO lead us to expect that primary HCHO may play an important role not only in ambient HCHO concentrations, but also in the radical and O₃ formation chemistry in the MCMA.

Employing a statistical regression analysis, Garcia et al. (2006) estimate the contribution of primary HCHO to the ambient HCHO concentration and conclude that a significant amount of HCHO observed in the MCMA is associated with primary HCHO. On a 24-h average, the contributions of primary, secondary and unaccounted sources (background) to ambient HCHO concentrations are 42, 37, and 21%, respectively, with the primary HCHO dominating at night and during early morning hours.

A detailed analysis of radical sources in the MCMA is presented by Volkamer et al. (2007) using a box model featuring the Master Chemical Mechanism constrained by an extensive array of measurements. These authors find that contributions of various radical sources (O₃ photolysis, HCHO photolysis, HONO photolysis, alkene ozonolysis and photolysis of other oxygenated VOCs (OVOC)) are fairly even in the MCMA. Using the HCHO partitioning results of Garcia et al. (2006), Volkamer et al. (2007) estimate that primary HCHO accounts for about 10% of the total radical production before 10:00 a.m., and contributes significantly to photochemical oxidants in mid-morning.

In this study, we employ the 3-D chemical transport model, the Comprehensive Air Quality Model with extensions (CAMx) (ENVIRON 2006), to investigate the impact of primary HCHO on the photochemistry in the MCMA, with important radical precursors and/or sources constrained by measurements from a local ambient air monitoring network and the MCMA-2003 campaign (Molina et al., 2007). The objective is to examine the influence of primary HCHO on the odd hydrogen radical budget and O₃ formation in the MCMA's atmosphere.

2 Methodology

2.1 Model and input

CAMx v4.40 with the SAPRC-99 gas phase chemical mechanism (Carter, 2000) was employed in this study. The model configuration and model domain are described in detail in Lei et al. (2007). Briefly, the model domain covers 52×52 grids with a grid resolution of 3 km centered in Mexico City; there are 15 vertical layers extending from the surface to about 5 km a.g.l with the bottom model layer interface of about 64 m a.g.l.

The model was driven by hourly meteorological output fields from the Pennsylvania State University/National Center for Atmospheric Research Mesoscale Model System (MM5, Grell et al., 1995), as described in de Foy et al. (2006). Gaseous emissions were estimated in two steps. First, spatially and temporally resolved and chemically speciated emission fields were constructed based on the official emissions inventory for the year 2002 in the MCMA (CAM, 2004) using the procedures and an updated database described in Lei et al. (2007); these emission fields are referred to as the initial emission estimates henceforth. Second, the initial emission estimates were then adjusted based on observations from the Mexico City Ambient Air Monitoring Network (RAMA) (SIMAT, 2003) and the MCMA-2003 field measurements (Lei et al., 2007). Model runs with varying emissions were carried out and simulated morning rush hour (06:00–09:00 a.m.) speciated VOC (as well as total odd nitrogen (NO_y) and CO) concentrations were compared with measurements until a good agreement was reached. Initial chemical concentrations and other boundary conditions were similar to those used in West et al. (2004) and were examined and adjusted using measurements taken at the boundary sites. Photolysis frequencies for clear sky under typical aerosol loading in the MCMA were precomputed using the TUV model (Madronich and Flocke, 1998) and were corrected for the effects of clouds using the algorithm of Chang et al. (1987). Details of the emission estimation and adjustments, as well as other model inputs, are described in Lei et al. (2007).

The simulation episode selected was 13–15 April 2003, for which Lei et al. (2007) has characterized MCMA ozone formation and its response to emission reduction. The base case (with primary HCHO) was similar to the reference case in Lei et al. (2007) except a new version of CAMx (v4.40 vs. v4.03) was used and the CMAQ scheme (Byun, 1999) was used to reconstruct the vertical diffusivity fields (k_v). Because of the changes to the vertical diffusion algorithm in the new version (Greg Yarwood, personal communication, 2007) and because CMAQ scheme overestimates k_v values (de Foy et al., 2007), the vertical diffusivity inputs were adjusted. By comparing k_v values with the results of the O'Brien scheme (1970) used in Lei et al. (2007) and by comparing concentrations of observed primary pollutants

(mainly CO, NO_y), a day-to-day varying scaling factor of 0.2–0.3 was applied uniformly (spatially) to the original k_v values.

The photochemical impact of primary HCHO is identified by turning on and off the HCHO emissions. As such, the effects of primary HCHO in HCHO concentration and photochemistry include not only the influence of the directly emitted HCHO (direct effect), but also the influence from the addition of photochemical HCHO produced through the enhanced photooxidation of hydrocarbons due to the addition of the primary HCHO (indirect effect). It should be noted that the effect (or impact, contribution, etc.) of primary and secondary HCHO defined here is different from those reported in Garcia et al. (2006) and Volkamer et al. (2007); we have included the indirect effect in the primary term, whereas the other articles include it in the secondary term.

2.2 Measurements and model constraints

One aspect of the complexity in urban photochemistry is the presence of various important radical sources in the urban atmosphere, which include O₃, O₃ precursors such as VOCs and NO_x, heterogeneous sources of HONO, and alkene-O₃ reactions in the case of the MCMA (Volkamer et al., 2007). A critical aspect for successfully quantifying the impact of primary HCHO depends on whether these radical sources in the urban areas, in addition to primary HCHO, are accurately captured in the simulation.

The emissions of CO, NO_x, and speciated VOCs used in the model were constrained by concurrent measurements, as described in Lei et al. (2007). Briefly, simulated CO, NO_x and O₃ were compared with RAMA observations; concentrations of alkanes and alkenes were compared with measurements obtained from canister sampling analyzed by Gas Chromatography/Flame Ionization Detection (GC/FID) at various sites during the MCMA-2003 campaign (Velasco et al., 2007); emissions of aromatics and formaldehyde were evaluated and adjusted by comparing with long-path Differential Optical Absorption Spectroscopy (DOAS) measurements at the campaign supersite CENICA (Volkamer et al., 2005). The DOAS data also included HONO concentrations. Olefin measurements using a Fast Isoprene Sensor (FIS) calibrated with a propylene standard at CENICA (Velasco et al., 2007) were also used in this study.

By comparing the simulated and the measured concentrations of HCHO and considering findings from previous studies on the mobile emission rate of HCHO in Mexico City (Kolb et al., 2004; Zavala et al., 2006), a factor of 7 was needed to scale the initial emission estimates of HCHO (from 2050 tons/yr to 13 920 tons/yr) in the MCMA in order to reproduce the measured HCHO concentrations. The factor of 7 is consistent with the findings from previous studies on the mobile emission rate of HCHO in the MCMA; both Kolb et al. (2004) and Zavala et al. (2006) show that on average the HCHO/CO₂ emission ratio in traffic exhaust plumes is a fac-

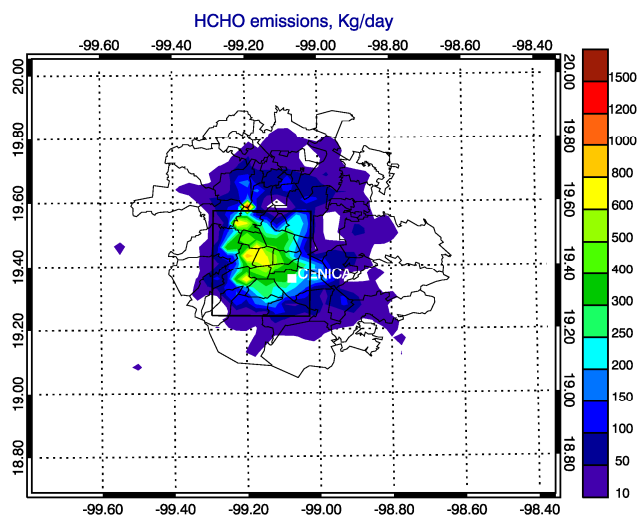


Fig. 1. Spatial distribution of HCHO emissions in the CAMx model domain. The daily emissions have been adjusted from 5615 kg/day (estimated in the 2002 EI) to 38 140 kg/day used in this study. The inner black rectangle designates the “urban region” used in the text. Also shown in the figure is the MCMA-2003 campaign supersite CENICA (white square).

tor of 7 higher in the MCMA than in the greater Boston area (2×10^{-4} vs. 3×10^{-5}), although the emission ratio varies with traffic. A map showing the spatial distribution of HCHO emissions is presented in Fig. 1, which also shows the locations of the CENICA supersite and the “urban region” – an area of $30 \times 36 \text{ km}^2$ encompassing the MCMA urban center and containing most of the RAMA monitoring stations. Figure 2 provides a comparison of simulated (using adjusted HCHO emissions) and observed HCHO concentrations at CENICA. The model well reproduced the magnitudes and diurnal variations of HCHO concentration using the adjusted HCHO emissions, except that it overpredicted HCHO concentrations by 40% during late afternoon and nighttime (from 16:00 p.m. to 05:00 a.m. next day). The overprediction is probably due to underestimated mixing, rather than overestimated emissions, since the model overestimation during this time window also occurred to CO and NO_y (Lei et al., 2007), and morning rush hour concentrations of HCHO were well reproduced. The HCHO detection limit of DOAS (1 ppb) may also contribute to the discrepancy since during this time window HCHO ambient concentrations were only a few ppbs (Fig. 2a). For the rest of the time, simulated HCHO were in excellent agreement with the measurements. As discussed later, the overprediction will affect the HCHO partitioning during late afternoon and nighttime, but should have little effect on the photochemistry, because the photochemical influence of primary HCHO is most significant in the morning and early afternoon.

In the MCMA, 85% of radical sources are produced from photochemical reactions (Volkamer et al. 2007), and light

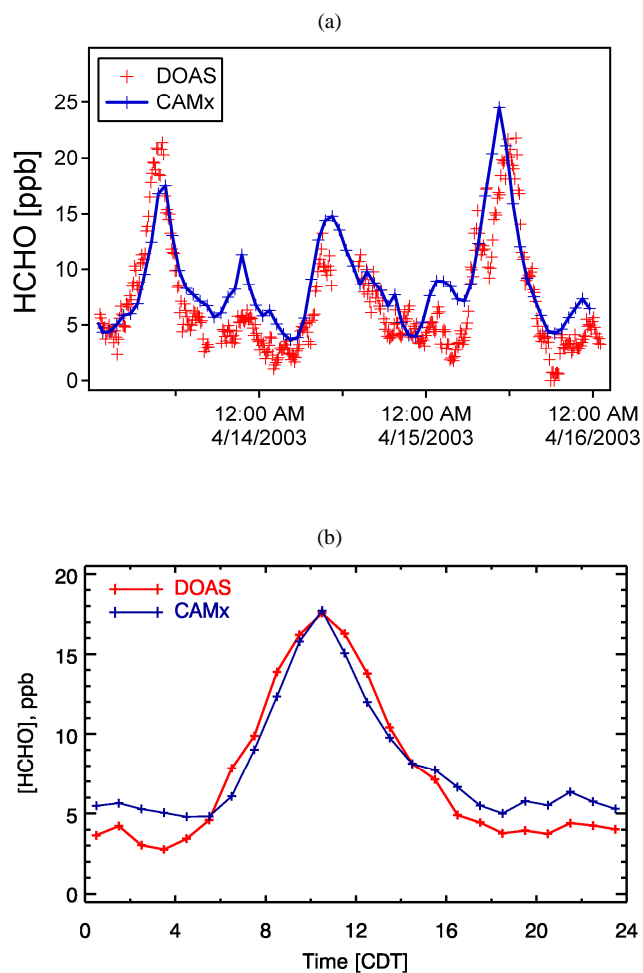


Fig. 2. Comparison of measured and simulated of (a) day-to-day variation and (b) episode-averaged (13–15 April) diurnal variation of surface HCHO concentrations at CENICA. DOAS measurements are denoted by the red crossed line, while simulated values are denoted by the blue crossed line.

absorbing organic aerosol reduces the amount of sunlight that reaches the ground, lowering radical production from photochemical sources such as HCHO, HONO, O₃, and OVOC photolysis (Barnard et al. 2008). To account for the effects of aerosol on the radiation using a first-order approximation, the aerosol optical depth at 340 nm was set to 0.8 for Mexico City in the TUV computation of photolysis frequency, conforming to the concurrent LIDAR aerosol extinction measurements reported during the 2003 campaign (Frey et al., 2004; Simeonov et al., 2004). In addition, a wavelength-independent value of aerosol single scattering albedo (SSA) in the urban area was set to 0.90. Barnard et al. (2008) report morning hour SSA values as low as 0.78 at 368 nm in the MCMA and SSA is wavelength dependent. We examined the simulated J-values in the urban area, and they were in good agreement with those measured by spectroradiometry during MCMA-2003 (Volkamer et al., 2005); for example,

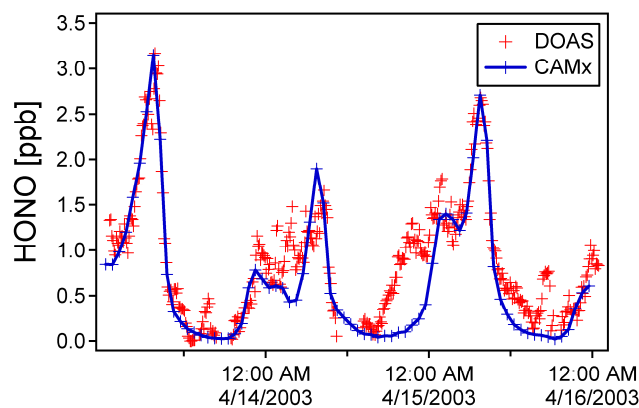


Fig. 3. Comparison of measured and simulated time series of surface HONO concentrations at CENICA. DOAS measurements are denoted by the red crosses, while simulated values are denoted by the blue crossed line.

JNO₂ was about 30 h⁻¹ at noon. It should be noted that although the near surface J-values were well simulated using a SSA value of 0.90 combined with other aerosol parameters, the assumption of smaller aerosol absorptivity would lead to an overestimate of J-values aloft, which in turn may lead to an overestimate of radical production.

Heterogeneous chemistry is a major source for nighttime HONO (Platt et al., 1980a; Kleffmann et al., 1998; Finlayson-Pitts et al., 2003; Stutz et al., 2004). However, since the formation mechanism is still not well understood, a direct emission of HONO was artificially included in the model as the surrogate for the heterogeneous source. The magnitude and temporal variation of the estimated HONO emissions were based on the NO emissions and were constrained by the DOAS HONO measurements; i.e., the temporal variation of HONO emissions were adjusted until the simulated HONO agreed with the measurements. A comparison of the simulated and observed HONO concentrations is shown in Fig. 3, which illustrates good agreement between the two.

Volkamer et al. (2007) find that alkene ozonolysis contributes significantly (~18%) to the radical production during morning hours in the MCMA. Oxidation of anthropogenic alkenes also plays a major role in daytime O₃ production in the MCMA (Tie et al., 2007) and in the US urban areas (Zhang et al., 2004), particularly in early morning where they dominate O₃ production. We did not have direct constraints on this radical source. Instead, this source was indirectly constrained by evaluating the simulations of O₃ and VOCs. Figure 4 shows the comparison of simulated and observed diurnal variations of surface concentrations of O₃ and alkenes at CENICA averaged over the simulation episode. The alkene (OLE) concentration shown in the figure is the propene-equivalent alkene concentration averaged over 14–15 April measured by the FIS. The OLE comparison

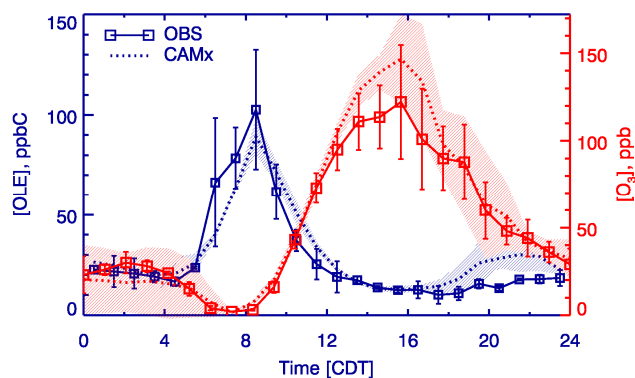


Fig. 4. Comparison of simulated and measured diurnal variations of surface alkene (OLE) in blue and O_3 concentrations in red at CENICA. Measurements in solid lines and simulations in dash lines. Error bars represent ± 1 standard deviations of measurements, and hatched areas represent ± 1 standard deviations of simulations.

did not include 13 April, which had anomalously high measured VOC concentrations. For the simulated values, concentrations of different olefin model species were weighted based on their FIS response factors and their contributions to the standard VOC mixture used in the SAPRC99 mechanism (Velasco et al., 2007). As shown in the figure, the simulated morning hour concentrations of O_3 and OLE agree well with the observations (within 1σ level). This suggests that the precursors of this radical source were well constrained, and demonstrates that the O_3 -OLE reaction is well represented in the model.

NO_3 -alkene reactions can also be an important source for nighttime peroxy radicals in both marine boundary layer and urban atmosphere (Platt et al., 1980b; Geyer et al., 2003; Lei et al., 2004). Volkamer et al. (2007) report the first direct evidence for the presence of up to 50 ppt NO_3 inside a megacity (Mexico City), and the relative contribution of this source strongly depends on altitude. Indeed Lei et al. (2004) find that NO_3 -alkene reactions contribute substantially to nighttime organic radicals in the residual layer in Houston, while their impact is negligible at ground level. In this study there were no constraints for this source.

3 Results and discussion

3.1 HCHO partitioning

Primary emissions, secondary photochemical formation by VOC oxidation, and background HCHO all contribute to the simulated concentrations of HCHO. The background HCHO in the model comes from transport outside the model domain, i.e., from the vertically-varying HCHO boundary condition which was set to 1.5 ppb at surface in this study. The very low VOC boundary conditions used in this study lead to a negligible contribution to the HCHO background through

Table 1. Contributions of emissions and secondary formation to ambient HCHO at CENICA.

Time	Primary (%)	Secondary (%)
00:00–06:00, 19:00–24:00	69	24
06:00–09:00(rush h)	80	8
11:00–18:00	32	67
daily	58 50 ⁺ 47*	37 44 ⁺

+ After correction for the 40% overprediction of HCHO concentration at night and early morning.

* Result of Garcia et al. (2006) when the primary HCHO indirect effect is included in the primary term.

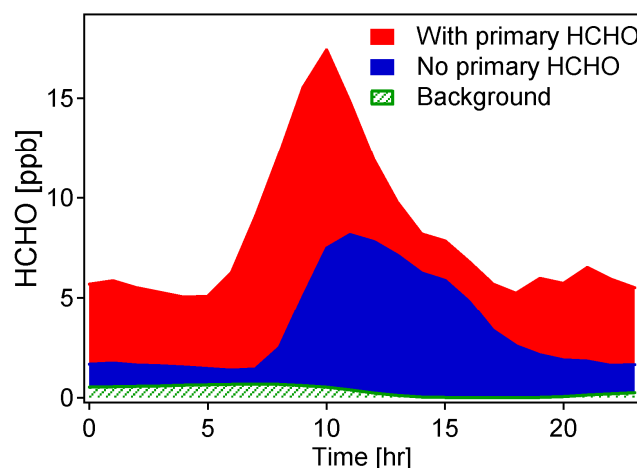


Fig. 5. Partitioning of HCHO to sources of primary (including the indirect effect), secondary and background at CENICA. The red area denotes the contribution of primary HCHO to the HCHO concentration, blue area denotes the contribution of secondary HCHO when there are no HCHO emissions, and the hatched green area represents background contributions.

photochemical oxidation. Figure 5 shows the simulated contributions of different sources to HCHO at CENICA. The contribution of primary HCHO was obtained as the difference between the run with primary emissions (base case) and the other control run without primary emissions. As pointed out earlier, the contribution of primary HCHO includes not only that of directly emitted HCHO, but also from the additional production of secondary HCHO through the enhanced photooxidation of hydrocarbons due to the addition of the emitted HCHO (indirect effect). The contribution of the boundary condition on the ambient HCHO concentrations was obtained by calculating the difference in HCHO concentrations from the control run above with a run where the HCHO boundary conditions were set to 0. The effect of the HCHO boundary condition is highest at nighttime but almost vanishes during the late morning to afternoon hours, because of the short chemical lifetime of HCHO in this period.

Table 2. Contribution of various radical sources to cumulative radical production in the urban region (unit:ppb). Note all radical terms here are not represented as the OH-equivalent quantities.

Time	OH _{new}				HO ₂ _{new}		RO ₂ _{new}	
	Total	O ₃ photolysis	O ₃ /OLE	HONO photolysis	Total	HCHO photolysis	Total	Aldehyde photolysis
07:00–09:00	2.2	0.1	0.2	2.1	2.1	0.5	0.9	0.1
09:00–13:00	5.7	3.8	1.5	0.4	20.9	7.5	9.0	1.4
13:00–16:00	7.7	6.7	1.0	0.0	12.8	4.8	5.3	1.0
16:00–20:00	2.8	1.3	1.4	0.0	4.6	1.2	3.3	0.2
06:00–18:00	17.6	11.7	3.3	2.4	39.2	13.9	17.1	2.6
Daily	19.5	11.8	5.1	2.4	42.5	14.1	24.4	2.7

Table 1 summarizes the contributions of primary and secondary (not including primary HCHO's indirect contribution) HCHO as a function of the time of the day. It can be seen from Fig. 5 and Table 1 that primary HCHO contributes significantly to the ambient HCHO concentration at night and early morning (~69%), and culminates at morning traffic rush hours (~80%). On the other hand, the contribution of secondary HCHO to the ambient HCHO becomes significant during the mid-morning and becomes dominant in the afternoon (~67%). On a 24-h basis, the contributions from primary, secondary and background sources are 58, 37 and 5%, respectively. A correction to the 40% model over-prediction (i.e., assuming a perfect agreement) of HCHO between 04:00 p.m. to 05:00 a.m. (see Sect. 2.2 and Fig. 2b) would modify the corresponding contributions to 50, 44, and 6%, respectively.

Garcia et al. (2006) estimate the partitioning of HCHO in the MCMA using a statistical regression analysis, and find that the contributions of the primary, secondary, and background (unaccounted) sources are 42, 38, and 21%, respectively, on a 24-h basis. Note the definitions of primary and secondary sources are different from Garcia et al. (2006), where the indirect effect of emitted HCHO is included in the secondary term. The model configuration we used did not allow us to quantify the indirect effect, but according to Volkamer et al. (2007), the indirect effect accounts for about 10% (upper limit) of the primary HCHO. Taking this into consideration, the contributions of primary, secondary and background sources to ambient HCHO in Garcia et al. (2006) would be 47, 32, and 21% respectively. These corrections bring the results of Garcia et al. (2006) and this study on the primary source contribution to the ambient HCHO concentration to an excellent agreement (within 3%), even though different methods and time coverage were employed. The definition of the background HCHO in Garcia et al. is significantly different because it represents the source unaccounted for by the tracers of CO (for emissions) and glyoxal (for secondary). A significant, if not major, portion of the background HCHO in Garcia et al. may consist of the photochemical secondary sources that are not correlated with gly-

oxal. Therefore direct comparisons for the secondary and background contributions in the two studies may not be appropriate.

3.2 Impact on radical budgets

Photolysis of HCHO directly produces HO₂. HCHO can also affect OH and RO₂ (organic peroxy radicals) through radical cycling processes. To estimate the overall effects of HCHO on radical concentrations and radical production rate, the term OH-equivalent (denoted as ΣOH, in unit of ppb/h) is used, which takes into account the propagation efficiency of the HO₂→OH and RO₂→OH conversions. Analogous to Sheehy et al. (2008), the conversion efficiencies of HO₂→OH, γ_{HO₂}, and RO₂→OH, γ_{RO₂}, are the portion of HO₂ and RO₂ participating in the radical propagation process that convert to OH during the propagation, respectively, which are defined as follows:

$$\gamma_{\text{HO}_2} = \frac{\text{HO}_2 \rightarrow \text{OH}}{\text{total HO}_2 \text{ reacted}} \quad (1)$$

$$\gamma_{\text{RO}_2} = \gamma_{\text{HO}_2} \frac{\text{RO}_2 \rightarrow \text{HO}_2}{\text{total RO}_2 \text{ reacted}} \quad (2)$$

Where HO₂→OH and RO₂→HO₂ are the propagation rates of HO₂ to OH and RO₂ to HO₂, respectively. As such,

$$\Sigma\text{OH}_{\text{new}} = \gamma_{\text{HO}_2} \times \text{HO}_{2\text{new}} + \gamma_{\text{RO}_2} \times \text{RO}_{2\text{new}} + \text{OH}_{\text{new}} \quad (3)$$

Where OH_{new}, HO₂_{new} and RO₂_{new} are the radical initiation rates of OH, HO₂ and RO₂ (in unit of ppb/h), respectively, i.e., the rates of radicals produced during the radical initiation process (also defined as new radical production rates). OH_{new} is primarily contributed directly by photolysis of O₃ and peroxides, alkene ozonolysis, and HONO photolysis. The HONO photolysis term here is actually the net OH production from the HONO photolysis minus the HO+NO→HONO reaction. HO₂_{new} is contributed by photolysis of carbonyls (such as aldehydes and ketones) and alkene-O₃ reactions. RO₂_{new} is contributed by photolysis of carbonyls and alkene-O₃/NO₃ reactions; radical production from net

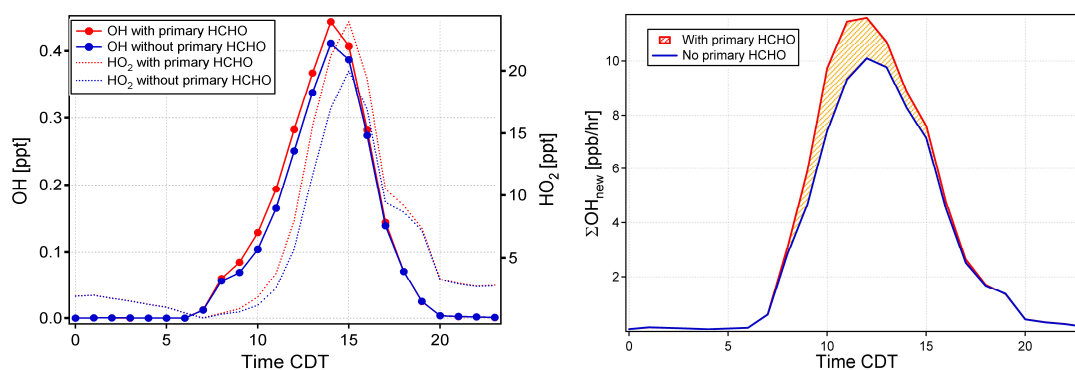


Fig. 6. Effects of primary HCHO on (left) radical concentrations and (right) radical production in the urban region averaged over 13–15 April 2003. In the left panel, solid lines represent the diurnal variations of OH concentrations, and dashed lines represent the diurnal variations of HO₂ concentrations. Inclusions of HCHO emissions are depicted in red, while exclusions of HCHO emissions are depicted in blue.

PAN chemical loss (when PAN thermal decomposition minus PAN formation is positive) is also included.

Table 2 shows the contribution of various radical sources to the radical production (initiation) as a function of time during the day (averaged over the episode) in the MCMA urban region. On a 24-h basis the HO₂ production is a major radical pool, followed by RO₂ and OH in that order; HCHO is an important HO₂ source, accounting for 33%. It is noted that during 07:00–09:00 a.m. although primary HCHO dominates ambient HCHO (~80%), its contribution to radicals (on the order of 80%*0.5 ppb=0.4 ppb) is less important than contributions from HONO photolysis (γ_{HO_2} is nearly one at this time) and O₃-OLE reactions (the major HO₂ production is from O₃-OLE reactions). This is consistent with the finding of Volkamer et al. (2007) which suggests that primary HCHO does not significantly influence the early morning photochemistry.

Figure 6 and Table 3 show the contributions of primary HCHO to surface radical concentrations and OH-equivalent radical initiation rates in the MCMA urban region. The percentage change for radical concentrations due to primary HCHO is obtained by calculating the enhancement (normalized by the base case) for each hour and then averaging over the hours of interest. The change in radical production is calculated by integrating the radical production rate over the time duration of interest and then computing the enhancement. It is evident that primary HCHO contributes significantly to the radical budget (starting from 08:00 a.m.), particularly in the morning and early afternoon hours in the MCMA urban area. For example, on the episode-average basis, the contribution of primary HCHO to OH concentration increases from 5% in the early morning to 16% in the late morning and early afternoon, peaking at about 20% at 10:00 a.m. It enhances peak OH concentration by 7%. For HO₂, the primary HCHO contribution increases from 10% at early morning to 31% at late morning and early afternoon, peaking at 34% at 10:00 a.m., and enhances peak concentra-

Table 3. Percentage increases of radical concentrations and initiation rates due to primary HCHO.

Time	$\Delta[\text{HO}_2]$, %	$\Delta[\text{OH}]$, %	$\Delta\Sigma\text{OH}_{\text{new}}$, %	
			+	*
07:00–09:00	10	5	6	5
09:00–13:00	31	16	17	17
13:00–16:00	20	7	8	8
16:00–20:00	10	5	6	5
06:00–18:00	18	8	12	10
Daily	11	5	12	5

+ Numbers are calculated by integrating the radical production rate over the time duration of interest then computing the enhancement.
* Numbers are obtained by calculating the enhancement for each hour then averaging over the hours of interest, same as for the calculation of the enhancement in the radical concentration.

tion by 16%. Note that its contribution to the HO₂ concentration remains important (20%) till late afternoon.

The contribution to the OH-equivalent primary radical production rate ($\Sigma\text{OH}_{\text{new}}$) is similar to that of OH concentration during daytimes, but the $\Sigma\text{OH}_{\text{new}}$ peaks much earlier. On the 24-h basis, primary HCHO enhances OH, HO₂ and $\Sigma\text{OH}_{\text{new}}$ by 5, 11 and 12%, respectively. The contributions of primary HCHO to daytime (06:00–18:00 LT) OH, HO₂, and $\Sigma\text{OH}_{\text{new}}$ are 8, 18, and 12%, respectively. If the enhancement for $\Sigma\text{OH}_{\text{new}}$ is calculated the way the radical concentration enhancement is computed, which is used in Volkamer et al. (2007), the average contribution of primary HCHO to daytime $\Sigma\text{OH}_{\text{new}}$ is 10%, as shown in Table 3. Unlike the contribution to ambient HCHO concentrations, which dominates during the night and early morning but decreases significantly after late morning, important contributions of primary HCHO to radical concentrations and production are evident after 09:00 a.m. and continue into early afternoon.

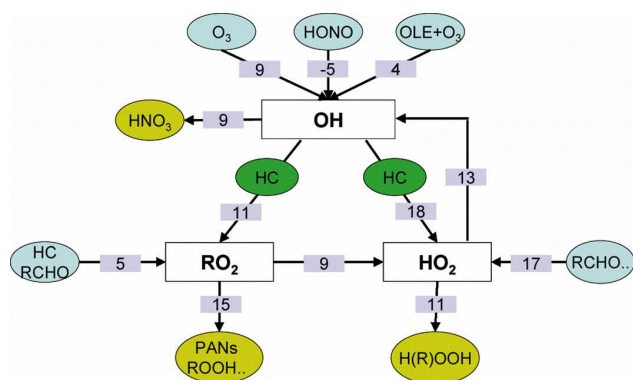


Fig. 7. Effects of primary HCHO on the radical flows in the urban region. The numbers on each arrow indicate the percentage change of radical chemical flux that follows each pathway accumulated over the entire simulation episode and over the urban region. RCHO includes HCHO. Radical sources are color-coded by light-blue and radical sinks by olive. H(R)OOH represents H_2O_2 and ROOH.

Using a box model with the Master Chemical Mechanism constrained by comprehensive measurements, Volkamer et al. (2007) quantified the radical sources in the MCMA. One of their findings is that primary HCHO (including the indirect effect) contributes up to 10% to $\Sigma\text{OH}_{\text{new}}$ in the morning and 7% for the 06:00–18:00 LT average (9 and 6% respectively if the indirect effect is excluded) at CENICA. Our results yield about 40% higher primary HCHO contributions; part of this discrepancy may be due to the assumption of aerosol's smaller absorptivity used in the TUV calculation, which would lead to an overestimate of UV radiation aloft, resulting in an overestimate of the relative contribution from primary HCHO, even though the total radical production would also be overestimated. Another issue is that in Volkamer et al. (2007) the HCHO source attributed to VOC oxidation was evaluated by setting all but one class of VOC or oxidant of interest to zero, which leads to an upper limit radical production from secondary HCHO. A third issue is that our results are derived from a 3-day simulation average, while Volkamer et al. (2007) use median constraints for the entire one month MCMA-2003 campaign. In addition, the spatial range is different (urban region vs. CENICA). Despite these differences, the agreement is still reasonable. The 06:00–18:00 LT average production rate of $\Sigma\text{OH}_{\text{new}}$ is 6.0 ppb/h, which is in excellent agreement with 6.1 ppb/h reported by Volkamer et al. (2007).

3.3 Impact on radical flow

Radicals are formed from the photolysis of O_3 , aldehydes, and other species; subsequently they undergo propagation reactions, and eventually are removed from the system by radical- NO_x reactions (e.g., formation of HNO_3 , PANs, and other organic nitrates) and radical-radical reactions (which

Table 4. Percentage increases of O_3 concentrations due to primary HCHO.

Time	$\Delta\text{P}(\text{O}_x)^+$, %	$\Delta[\text{O}_3]^*$, %
07:00–09:00	10	6
09:00–13:00	18	17
13:00–16:00	8	8
16:00–20:00	8	3
Daily	13	6
Peak O_3	12	8

+ Calculated as integrating over hours first then computing the enhancement.

* Calculated as computing the hourly enhancement first then doing average over the hours.

lead to the formation of peroxides). The addition of primary HCHO to the system modifies these radical initiation-propagation-removal processes (radical flow). By tracking the influences of primary HCHO on the radical flow, we can identify the roles of primary HCHO in the radical chemistry.

The effect of primary HCHO on the radical flow in the MCMA is illustrated in Fig. 7, where the number on each arrow indicate the percentage change of radical chemical flux that follows each pathway accumulated over the entire simulation episode and over the urban region. As primary HCHO is added, new OH radicals produced from O_3 photolysis and alkene ozonolysis increase by 9 and 4%, respectively, as a result of increases in O_3 concentration. In contrast, the radical formation from the net HONO source (HONO photolysis minus $\text{OH}+\text{NO}\rightarrow\text{HONO}$) decreases by 5%. This decrease is due to the larger magnitude of the $\text{OH}+\text{NO}\rightarrow\text{HONO}$ reaction over the photolysis in the early morning (07:00–09:00 a.m.) when primary HCHO is included, which leads to a marginal increase in HONO concentration (not shown), consistent with the HONO source apportionment during morning hours reported by Volkamer et al. (2007). Initiation rates of HO_2 and RO_2 are increased by 17 and 5%, respectively, which are attributed to the increase in HCHO and other species directly from emissions and indirectly from the additional VOC photooxidation. The interconversions between different radicals increase significantly (9–18%). The radical removal processes are also enhanced: 9% for HNO_3 , 15% for organic nitrates, and 11% for peroxides. The changes in the interconversion rates are basically equal to the changes in radical removal rates, implying that the catalytic efficiency of radicals in O_3 formation remains largely unchanged.

In summary, primary HCHO significantly enhances radical initiation, propagation and termination. The impact of primary HCHO on radical budgets described in the previous section is the ultimate consequence of the impact of primary HCHO on the radical flows.

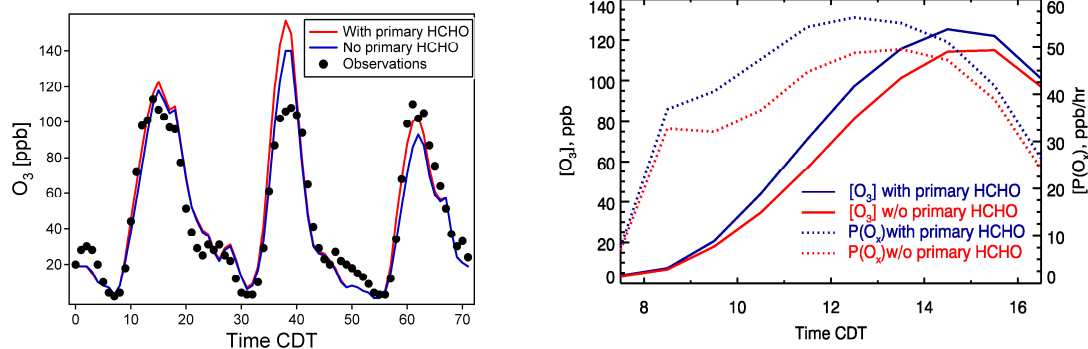


Fig. 8. Effects of primary HCHO on surface O_3 production and concentration in the urban region. RAMA measurements in black dots, simulations with primary HCHO in red lines, and simulations without primary HCHO in blue lines. The left panel shows the O_3 time series over the whole simulation episode, while the right panel focuses on the morning and afternoon hour O_3 production (dotted lines) and concentration (solid lines) averaged over the episode.

3.4 Impact on O_3 formation

Due to the changes in radical budgets, primary HCHO is expected to influence O_3 photochemical formation in the MCMA. Figure 8 and Table 4 shows the time series of simulated surface O_3 concentration and photochemical production rates $P(O_x)$ ($O_x=O_3+NO_2$) averaged over the urban region, with and without primary HCHO emissions. The enhancement of ambient O_3 concentration by primary HCHO starts to emerge during the morning rush hours with values up to 6%. This increase should be an upper limit since the early morning hour HCHO concentration was overestimated as mentioned earlier. Primary HCHO increases the surface peak O_3 concentration by 4% to 10% (corresponding to changes of 4–17 ppb), varying from day to day, but 8% on average. Moreover, the increase of peak O_3 concentration due to the primary HCHO is more evident and magnified at locations producing high O_3 levels (not shown). The significance of the 8% contribution of primary HCHO to O_3 concentration can be gauged by considering that a 50% reduction in emissions of both VOC and NO_x leads to the decrease of urban surface peak O_3 concentrations by 19% (Lei et al., 2007). The most evident impact of primary HCHO on surface O_3 occurs at mid morning and during the early afternoon (09:00–13:00 LT), where primary HCHO leads to an increase of O_3 concentration by 17% on average (over the episode) in the urban region. After 16:00, the contribution of primary HCHO decreases rapidly. As in the case of radicals, the significant impact of primary HCHO on O_3 continues into the late afternoon. The effect of primary HCHO on O_3 production rates is similar to that of O_3 concentration (the higher increase of $P(O_x)$ at photochemically-dark times, when changes in both O_3 and $P(O_x)$ are nearly zero, is mainly due to the difference in the enhancement calculation procedure). Nevertheless the effect of primary HCHO on $P(O_x)$ starts earlier than the effect on O_3 concentration (note the difference at 07:00–09:00 a.m.). The rapid increase

in $P(O_x)$ from 08:00 to 09:00 a.m. in Fig. 8 is due to the HONO radical source.

Primary HCHO affects not only the magnitude of O_3 concentration and photochemical production rates, but also their timing. As shown in Fig. 8, although there are flat and broad peaks in both $P(O_x)$ and O_3 concentration (model output frequency is 1-h), it appears that the primary HCHO causes $P(O_x)$ to peak half hour earlier ($\sim 12:30$ LT vs. 13:00 LT). It is expected that the timing shift would stand out with a higher output frequency. The peak $P(O_x)$ time shift and greater enhancement of $P(O_x)$ at earlier time stated above lead O_3 concentrations to peak about a half hour earlier too ($\sim 14:30$ LT vs. 15:00 LT), even though the peak O_3 concentration is flat and broad. The timing shift in peak O_3 concentration due to primary HCHO suggests that O_3 production is VOC limited in the MCMA source area. Lei et al. (2007, 2008) found that a 50% reduction in NO_x emissions or a 50% increase in VOC emissions leads O_3 concentration to peak one hour earlier. The timing change together with the influence on peak O_3 concentration demonstrates the importance of primary HCHO in O_3 formation in the MCMA.

Finally, it should be noted that our current understanding of the radical production process is incomplete (Volkamer et al., 2007; Sheehy et al., 2008). Radical production fluxes during early morning are uncertain by about a factor of two. Due to the lack of understanding of the underlying physicochemical processes that are causing this mismatch, our discussion of primary HCHO does not attempt to account for this effect. Any additional radical flux from sources that are currently unidentified will reduce the estimates of the contributions of primary HCHO to the radical budget, as well as ozone formation.

It has been reported that the reaction of electronically-excited NO_2 (NO_2^*) with water vapor could be a potentially important OH radical source at high solar zenith angles in the troposphere (Li et al., 2008), and inclusion of this source

could improve the estimation of radical budget and O₃ production in the MCMA (Spencer et al., 2009). However, the significance of this source for Los Angeles O₃ production is disputed (Wennberg and Dabdub, 2008). Our preliminary CTM simulation results using the rate constant reported by Li et al. (2008), which are in preparation for publication, show that the NO₂^{*}+H₂O radical source is negligible to O₃ production in the polluted atmosphere over the MCMA where photochemistry is characterized by complicated radical sources.

4 Conclusions

A distinctive feature of the Mexico City metropolitan Area is the high emissions of formaldehyde from anthropogenic sources, in particular from vehicle exhaust. By employing the CAMx chemical transport model for a case run on 13–15 April 2003, the impact of primary HCHO on the ambient HCHO concentration, radical budgets and O₃ formation in the MCMA was examined. Important radical sources, including HCHO, heterogeneous HONO, alkene ozonolysis and other radical precursors, were constrained by measurements from the MCMA-2003 campaign and MCMA's routine ambient air monitoring network.

Primary HCHO contributes significantly to the ambient HCHO concentration (up to 50% on daily basis); primary HCHO dominates ambient HCHO at night and in the morning (accounting for up to 80%), and decreases to 32% or less in the afternoon.

Primary HCHO contributes significantly to MCMA's radical budget, including OH and HO₂ concentrations, and the OH-equivalent radical initiation rate $\Sigma\text{OH}_{\text{new}}$, particularly in the morning and early afternoon hours in the urban area. The contribution of primary HCHO to radical formation becomes significant in the early morning, culminates at 09:00–11:00 a.m., and continues throughout the early afternoon. It enhances daily OH, HO₂ and total radical initiation rate $\Sigma\text{OH}_{\text{new}}$ by 8, 11 and 5%, respectively, while enhancing daytime OH, HO₂, and $\Sigma\text{OH}_{\text{new}}$ by 8, 18, and 12%, respectively. Through its direct effect (radical formation from primary HCHO itself) and indirect effect (radical formation from enhanced VOC oxidations due to primary HCHO), primary HCHO influences the radical initiation, propagation, and termination, which leads to the increase of OH, HO₂ and $\Sigma\text{OH}_{\text{new}}$.

Primary HCHO also significantly influences surface O₃ formation in the MCMA. Similar to the case for the radical budget, the influence and enhancement of O₃ formation by primary HCHO kicks-off in the early morning, peaks at mid-morning, and continues through the early afternoon. It not only increases the morning–early afternoon hour surface O₃ concentration (up to 19%), but also increases surface peak O₃ concentration (8% on average). Moreover, the increase and enhancement of peak O₃ concentration due to the primary HCHO is more evident and magnified at locations with

high O₃ levels. Primary HCHO also tends to make O₃ both production rates and concentration peak half an hour earlier.

The conclusions presented in this study are based on one case study, which covers a three-day episode, and the model constraints on important radical precursors, such as HCHO, HONO, alkene, and O₃ heavily rely on measurements at a specific location (CENICA). More studies are needed in order to comprehensively characterize the photochemical impact of primary HCHO with more spatial coverage and broader spatial coverage to improve measurement constraints. In particular it is vital to have an accurate estimate of HCHO emissions, which requires larger measurement datasets. We plan to further our modeling studies using the much larger and comprehensive datasets from the 2006 MILAGRO Campaign (Molina et al., 2008). In addition, it will be interesting to examine the influence of primary HCHO on the O₃-NO_x-VOC sensitivity chemistry.

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